

**Risk Assessment for Exposure to Ambient Air in the Vicinity
of the BKK Landfill in West Covina, 1978-1985**

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1. Introduction

The BKK landfill in West Covina has been in operation as a repository for toxic waste since 1976. Some of the residents in the vicinity of the landfill have complained that substances in the landfill are causing them to suffer an increased risk of disease and have asked that the landfill be closed. In 1983, the California Department of Health Services (DHS) *et al.* (DHS 1983) issued a report which contained measurements of the ambient air concentrations of volatile organic compounds, which the Department considered to be potential human carcinogens, and also reported corresponding measurements at a control site located 9 miles away. These measurements were used as a basis for calculation of the excess cancer risk associated with residence near the BKK landfill (DHS 1983, attached as Appendix 1). A review of the DHS 1983 report indicated that it was not possible to reproduce this calculation since no information was given with regard to what animal cancer tests were used or how certain factors, such as meteorological conditions, were accounted for. We were therefore asked to perform our own calculation of the excess cancer risk that may be incurred by residence in the vicinity of the BKK landfill during a 7-year period between 1978 and 1985. Our analysis uses data from exposure monitoring and from animal experiments along with a variety of methods and assumptions that are used by the U.S. Environmental Protection Agency (EPA), plus several new methods that have been developed recently. Section 2 describes these methods, Section 3 reports the results of the calculations we made, and Section 4 discusses these results.

2. Methods

In order to predict increased lifetime cancer risk, it is necessary in this case to calculate both a dose and a carcinogenic potency for each of the compounds to which individuals are purportedly exposed around the BKK landfill.

Dose is, of course, calculated from the available exposure information. Carcinogenic potencies are derived from the selected animal bioassay data. Our assumptions and methods regarding the calculation of exposure and potency are described below.

2.1. Methods & Assumptions Regarding Exposure Assessment

In the absence of detailed information on exposure in the vicinity of the BKK landfill, a number of assumptions were made regarding how to interpret the concentration data available to us as well as how to estimate duration of exposure, dilution of ambient contaminant concentration, absorption of inhaled compounds and, finally, the impact of seasonal concentration variability. These assumptions are described below.

2.1.1. Concentration Data

It is assumed that the mean ambient air concentrations (i.e., the mean value of the 2 [1 in the case of benzene] compound-specific mean concentrations) reported in DHS (1983) equal the actual time-weighted average (TWA) concentrations for the seven compounds measured, over the 7-year period of interest. These data were collected at 4 monitoring stations (Stations A, B, D, and F) located on the perimeter of the BKK landfill, as well as at a control monitoring station located 9 miles away (Station C). It is also assumed that the relative ratios of the latter values accurately reflect the relative contribution of each compound to total exposure over this period, for each monitoring station.

For the purpose of our analysis, we have further assumed that the TWA concentrations at the stations on the perimeter of the BKK landfill and those at the control station are in fact different. That is, we assume that the DHS-reported differences between the mean concentrations for Stations A, B, D and F, on the one hand, and Station C, on the other, are not due purely to chance.

If a proper statistical analysis of the the actual concentration measurement data indicate that the latter differences are in fact due to chance, then the excess risk is of course 0 without reference to any further assumptions. Since the raw concentration data were unavailable to us, we are unable to evaluate the likelihood of this possibility.

Increased concentrations of the 7 contaminants measured by DHS in the BKK vicinity above the corresponding control station values were calculated for stations A, B, D and F in terms of mg/kg/day. These values are given in Appendix 2. Note that for Station F, the mean of the measured values is less than the corresponding control station value for benzene and chloroform.

2.1.2. Exposure Duration

Here we assume that exposed residents are present in the vicinity of the BKK landfill for an average of 15 hours per day, for 365 days per year over a 7-year period.

2.1.3. Concentration Dilution

We assume that concentrations reported on the perimeter of the landfill are dispersed in the ambient air to some degree before being inhaled by the average resident, and that this process reduces the average concentration to which residents are exposed by a factor of 2. We further assume that no additional dilution takes place (e.g., due to filtration by air conditioning).

2.1.4. Absorption

We assume residents' average respiratory intake rate is 20 m³/day per 70 kg body weight (i.e., 0.2857 m³/kg/day). In their analysis of cancer risk due to household exposure of chlorinated hydrocarbons, such as those considered here, Cothorn et al. (1984) make the assumption that 50% of the amount of such compounds contained in inhaled air is absorbed into to blood and is

distributed via circulation, and the balanced is exhaled back into the atmosphere. We follow the Cothern *et al.* assumption of 50% absorption (i.e., 50% retention) of contaminants contained in respired air.

2.1.5. Seasonal Variation in Concentration

It is possible that the release of volatile organic compounds, such as those considered here, from the surface of the BKK landfill may be enhanced with increasing temperature, resulting in increased ambient concentration during the warmer months of the year. We note that the DHS concentration measurements were made between July and October of 1982. Since these months are the warmest of the year in southern California, these measurements may overestimate TWA concentrations at other times of the year. In the absence of other measurements, however, we make the conservative assumption that the DHS measurements accurately reflect TWA concentrations over the multi-year period of interest without seasonal adjustment.

2.2. Methods & Assumptions Regarding Carcinogen Potency Assessment

Since the mid-1960s, the primary basis for judging both carcinogenicity (i.e., carcinogenic potential) *per se* and the carcinogenic potency of particular chemicals to humans, in the absence of adequate epidemiological data, has been data from long-term carcinogen bioassays using laboratory animals. By this means, an assessment of human carcinogenic potency of a compound at very low environmental dose levels begins with an assessment of tumorigenic potency at much higher experimental dose levels in the test species used. This is followed first by an extrapolation of expected response in that species from these experimental dose levels down to the environmental levels of interest, then by an extrapolation of the latter predicted response, in the test species used, to a response in a target human population.

The validity of risk predictions calculated in accordance with the above procedure, of course, is conditional on the validity of key assumptions or "inference bridges" that must be used to bridge fundamental gaps that currently exist in basic scientific knowledge concerning the mechanism of carcinogenesis, the extrapolation of high-dose observations to low-dose predictions for a given species, and the equivalence of tumorigenic doses between different species. Environmental cancer risk assessment cannot proceed without the use of these assumptions, so that much effort has gone into studying how they may be standardized for use in a regulatory context (NAS 1983). But this does not mean that the uncertainty associated with these assumptions is in any way irrelevant or negligible, or even that their use will necessarily lead to risk predictions that are consistent with the "best guesses" of a majority of informed scientists/experts.

In certain components of our potency assessment we have made use of some of the assumptions and procedures that have become standard in the context of regulatory carcinogen risk assessment. But in other components of our analysis, we have made use of a range of plausible assumptions (rather than a single conservative assumption) in order to lend a greater degree of objectivity to the analysis. The validity of the final result, however, remains conditional on the validity of the residual set of inference bridges we have used for this analysis.

The following subsections discuss the specific methods and assumptions we use for carcinogen potency assessment in the present analysis.

2.2.1. Selection of Animal Data

The animal data we used for each compound were derived from EPA Health Assessment Documents when these were available. In the absence of such documents (for the cases of the 2 known human carcinogens vinyl chloride and

benzene), data from the scientific literature were selected in general accordance with EPA's preferred selection criteria. The latter criteria are *conservative* in the sense that efforts are made to *err on the side of safety* by generally focusing on data obtained for the most sensitive species/strain/sex tested while ignoring the weight of evidence for non-carcinogenicity or for less potent carcinogenicity presented by any remaining data (EPA 1984a).

Below we describe the specific data sources we used in our potency assessment. The actual dose-response input data used and related information appears in Appendix 2.

We note that in the health assessment documents for the five compounds for which EPA has produced such documents to date, animal dose levels used to extrapolate tumorigenic response for given bioassay data sets represent lifetime TWA *metabolized* doses. In these cases, and in the case of vinyl chloride discussed below, dose levels were (linearly) transformed back into an applied dose level that corresponds with expected applied low-level dose, fully in accordance with the pharmacokinetic data relied on in each case for determining metabolized dose. The latter transformation obviates the need to readjust calculated potencies that are based directly on a metabolized dose, rather than on an applied dose such as that calculated for BKK-vicinity residents. (See, however, the discussion of vinylidene chloride bioassay data below.)

We note finally that in certain cases several animal bioassay data sets are discussed and analyzed in EPA health assessment documents, whereas only a subset of these are actually used to calculate potency values recommended for use (albeit, for use in a *regulatory* context) by EPA. We have made use of only data sets relied on by EPA for final potency assessment in these cases.

1. Vinyl Chloride (VCL). No EPA health assessment document exists for VCL at this time. For potency assessment, we used the data of Maltoni and LeFem-

ine (1975) on the induction of zymbal gland tumors, nephroblastomas and liver angiosarcomas in female Sprague-Dawley rats exposed to VCL by inhalation. The experimental exposures in this study (0, 50, 250, 500, 2500, 6000, and 10000 ppm) were converted to lifetime TWA exposures in mg/kg/day using standard assumptions employed by EPA (Anderson *et al.* 1983, EPA 1984a). In addition, the resulting animal TWA doses were adjusted to reflect non-linear metabolism of VCL by rats at these concentrations, in accordance with the pharmacokinetic analysis done by Gehring *et al.* (1978).

2. Perchloroethylene (PCE). Bioassay data on PCE liver carcinogenicity in B6C3F1 mice have been used by EPA to extrapolate carcinogenicity at low doses (EPA 1984b). More recent bioassay data have been produced by the National Toxicology Program which may also be used to evaluate the carcinogenic potential of PCE (NTP 1985). See discussion in Section 2.2.2 regarding our omission of PCE data.

3. Trichloroethylene (TCE). Bioassay data on TCE liver carcinogenicity in B6C3F1 mice have been used by EPA to extrapolate carcinogenicity at low doses (EPA 1985a). See discussion in Section 2.2.2 regarding our omission of TCE data.

4. Vinylidene Chloride (VDCL). We make use of bioassay data on VDCL's capacity to induce kidney adenocarcinomas in male Swiss mice reviewed in EPA (1985b). EPA's derivation of dose data for this bioassay relies on experimental information regarding VDCL metabolism in rats and mice. For potency assessment, EPA assumes 100% metabolism of VDCL, and so uses calculated metabolized dose directly in its potency assessment. However, the metabolic data relied on by EPA for VDCL potency assessment show that 0.035 kg mice and 0.25 kg rats exposed to 10 ppm VDCL for 6 hours metabolize 0.186 mg and 0.723 mg VDCL, respectively. But using EPA's recommended method to calculate mouse

respiration rates in the context of cancer risk assessment (EPA 1984a), the corresponding inhaled amounts (which equal the respective applied doses only if 100% absorption is assumed—see discussion above) are

$$\text{Mouse: } \left[10 \text{ ppm} \times \frac{97 \text{ g/M}}{24.45 \text{ L/M}} \right] \times \left[0.0345 \left(\frac{0.035}{0.025} \right)^{2/3} \text{ m}^3/\text{day} \times \frac{6 \text{ hr}}{24 \text{ hr/day}} \right] = 0.429$$

$$\text{Rat: } \left[10 \text{ ppm} \times \frac{97 \text{ g/M}}{24.45 \text{ L/M}} \right] \times \left[0.105 \left(\frac{0.25}{0.113} \right)^{2/3} \text{ m}^3/\text{day} \times \frac{6 \text{ hr}}{24 \text{ hr/day}} \right] = 1.769$$

That is, 43% and 45% of applied dose is metabolized by mice and rats exposed to 10 ppm VDCL for 6 hours, respectively. Since metabolic saturation in these species occurs at a much greater dose level than 10 ppm, it may be argued that the appropriate applied dose data to use for mice in analyzing the VDCL bioassay data can be linearly extrapolated down to the ppb range of interest, i.e., that the dose data should be increased by a factor of $\frac{1}{0.43} = 2.3$ over the figures used by EPA. We did not correct for this source of potential conservatism in deriving appropriate dose data to use in the analysis of the VDCL bioassay data, but rather used EPA's dose data directly.

5. 1,2-Dichloroethane (DCE). We make use of bioassay data on DCE's capacity to induce hemangiosarcomas in male Osborne-Mendel rats and hepatocellular carcinomas in male B6C3F1 mice reviewed in EPA (1985c).

6. Chloroform (CLFM). We make use of bioassay data on CLFM's capacity to induce hepatocellular carcinomas in male and female B6C3F1 mice reviewed in EPA (1985d).

7. Benzene (BNZ). No EPA health assessment document exists for BNZ at this time. For potency assessment, we used the data of Maltoni (1983) on zymal gland tumor induction in female Sprague-Dawley rats exposed to BNZ by ingestion. Metabolism was not taken into account regarding the dose levels in

this data set.

2.2.2. Dose-Response Extrapolation

The extrapolation of observed response in animals at experimental doses to predicted response at low doses can be strongly influenced by the model used to carry out this extrapolation. This is clearly illustrated in Cothorn *et al.* (1984). EPA uses the "linearized multistage" model to extrapolate lifetime human cancer risk at very low dose levels based on a given set of animal carcinogenicity bioassay data (Anderson *et al.* 1983, EPA 1984a). This method yields an upper (one-tailed) asymptotic 95% confidence limit of low-dose potency for a given input data set. The uncertainty referred to by this confidence limit is that due only to estimation error in fitting the multistage model, where this model is simply assumed to be correct for the purpose of doing a quantitative potency analysis. This dose-response model is such that risk is approximately equal to dose multiplied by potency at very low dose levels, such as the levels we consider here.

In addition, low-dose potency and risk extrapolations generated by the multistage model are conditional on the assumption that the compounds whose potencies are extrapolated are indeed human carcinogens at low doses. While in a regulatory context, a number of *conservative, safety-oriented* assumptions have been adopted to enable analysts to predict the cancer risks which may possibly be associated with exposure to certain compounds (NAS 1983, EPA 1984a), these assumptions do not necessarily provide the *most likely* or *most reasonable* estimate of risk given current scientific knowledge. Accordingly, we do not feel that the compounds PCE and TCE should be included in our assessment of the *most likely* risk associated with exposure to emissions from the BKK landfill. This is because both of these compounds are unlikely to be actual human carcinogens at the low doses considered in this analysis. Both

compounds produce tumors in the livers of B6C3F1 mice and do not do so in the livers of rats. The B6C3F1 mouse is extraordinarily susceptible to liver cancer, possibly because of the presence of an oncogene in its liver cell DNA (Fox and Watantabe 1985). Therefore, any agency which produces toxicity in the livers of B6C3F1 mice is more than likely to produce liver tumors by stimulating cell proliferation rather than by causing genetic damage. Recent studies by Mirsalis *et al.* (1985) using both TCE and PCE support this proposition, i.e., that TCE and PCE produce liver cancer in B6C3F1 mice by an epigenetic toxic mechanism. Furthermore, mice do not metabolize TCE in the same manner as humans and do so at a much faster rate than either rats or humans. Since neither TCE nor PCE nor both will produce liver toxicity in humans at the low doses being considered here, they are extremely unlikely to produce liver cancer in the human population concerned.

When more than one data set is used to calculate carcinogenic potency, EPA uses the geometric mean of the calculated asymptotic upper confidence limits based on those data sets as an overall estimate of potency. The latter procedure, for instance, was used in EPA's recent calculation of TCE's carcinogenic potency based on 4 bioassay data sets (2 sexes in 2 different studies) (EPA 1984b).

For the purpose of deriving a lifetime tumorigenic potency assessment which takes a more objective approach to the uncertainties arising in dose-response extrapolation, we have used an alternative to EPA's method, called the Monte Carlo Potency Analysis (MCPA) method, to assess carcinogenic potency under different assumptions, such as regarding which data set to use (or how to do inter-species dose extrapolation, discussed below). The MCPA method uses the same multistage model used by the EPA method. However, the MCPA method relies on Monte Carlo simulation, rather than the more conservative

and approximate asymptotic approach used in the EPA method, and it produces *complete distributions* for potency estimation error, rather than the single upper confidence value for potency produced by the EPA method (Bogen 1986). Our application of the MCPA method to the above-specified dose-response input data was carried out using a version of the program GLOBAL79 used for the analysis of dichotomous tumor response data (Crump and Watson 1979) which was modified to generate complete Monte Carlo distributions of site-specific as well as composite tumorigenic potency based on that data (Bogen 1986).

Time-to-tumor data were not used in our analysis of carcinogenic potency, although it would be possible to do so in the context of our analytical framework using a time-dependent multistage model (Crump and Howe 1984). EPA potency assessments have occasionally used time-to-tumor data with such a model when these data are available and when early mortality in a bioassay is prevalent to the extent that it may lead to distorted estimates of potency. Potency values thus derived tend to yield slightly greater values for potency than are derived using a multistage model that does not incorporate time-to-tumor information. For example, for compounds considered herein that are addressed by EPA health assessment documents, upper confidence limit potency values derived using time-to-tumor information are higher than those derived without this information by a factor of approximately 2 or less. However, the incorporation of time-to-tumor information into potency assessment requires several additional assumptions, such as the selection of an appropriate exponent of time (which corresponds to an assumed number of stages in the multistage carcinogenic process) to use in the more complicated time-to-tumor model (Crump and Howe 1984). This introduces yet additional sources of uncertainty into the analysis. Since the time-dependent and time-independent models yield similar potency estimates in the cases we consider for which a comparison was undertaken by EPA in its health assessment documents, we

have relied only on the time-independent multistage model for our analysis.

Depending on the nature of a particular data set, the question may arise regarding how or whether to combine experimental incidence data for each significant tumor type observed in a given bioassay into data amenable to a quantitative assessment of increased *composite* risk, where here the term "composite" is used to refer to the potential or actual occurrence of *any one or more* types of tumor from among a prespecified set of types. EPA has used a tumor counting procedure that consequently has become widely applied in the context of environmental carcinogen risk assessment. According to EPA, if data exist for two or more significant tumor types in the same study, the number of animals with *at least one* of these types is used as incidence data for quantitative assessment of composite risk when bioassay data must be relied on for this purpose (Anderson *et al.* 1983, EPA 1984a). This procedure has been criticized in that it may lead to results which are inconsistent with the underlying tumor incidence data used for a given potency assessment, and an alternative procedure, "composite potency analysis," is available which always yields consistent potency estimates based on tumor incidence data segregated by tumor type (Bogen 1986). We have made use of composite potency analysis in the context of our assessment of VCL carcinogenic potency, in which case 3 different tumor types were observed to occur with significantly increased incidence in rats.

2.2.3. Inter-Species Dose-Equivalence Extrapolation

Equivalent doses between species may be calculated on the basis of dose per body weight or dose per body surface area. At present it is not known which inter-species dose extrapolation assumption better reflects reality in the context of extrapolating tumor response data in animals to anticipated response in humans, and existing data do not rule out either approach (Hogan

and Hoel 1982). When data from a bioassay using mice, e.g., are used for dose-response extrapolation, the choice of which inter-species dose extrapolation method to use alters the resulting potency calculation by a factor of more than 10. To provide an objective approach to the issue of inter-species dose extrapolation, we used a Monte Carlo approach to combine the potency distributions based on the body weight and surface area methods, under the assumption that each method is equally likely to reflect reality.

2.3. Methods & Assumptions Regarding Risk Calculations

To calculate approximate increased low-dose risk to BKK-vicinity residents, calculated excess doses in mg/kg/day for each of the compounds VCL, VDCL, DCE, BNZ and CLFM at each station were multiplied by the factor

$$\frac{7}{70} \times \frac{15}{24} \times 0.50 \times 0.50 = 0.015625$$

in accordance with the assumptions we made regarding exposure duration, concentration dilution and absorption. These excess doses were then multiplied by the corresponding calculated MCPA potency distributions to yield approximate compound-specific risk distributions for each of the stations A, B, D, and F, in accordance with the approximate linearity of the multistage dose-response model at very low doses. The latter station-specific risk distributions were then summed, using a Monte Carlo procedure, to yield distributions of total increased risk due to all 5 compounds for each station, under the assumption that the carcinogenic potency of the mixture of compounds is in all cases equal to the simple sum of the individual compound-specific potencies. That is, we assumed that there is neither positive nor negative synergism among the carcinogenic potencies of the mixture of compounds considered.

An overall BKK-vicinity risk distribution was then derived by taking a weighted average of the 4 calculated station-specific risk distributions, again

using a Monte Carlo procedure, where the station-specific weights were all assumed to be equal (i.e., all equal to 0.25, resulting in a simple average of the 4 distributions). Finally, from this average distribution of total increased risk, the median value (i.e., the 50th percentile value) was selected to represent a risk value which is as likely as not to overestimate the true level of average increased individual risk which BKK-vicinity residents have incurred over the 7-year period of interest, conditional on the set of assumptions we made in our risk analysis.

3. Risk Analysis Results

Our calculated MCPA potency distributions for each of the compounds VCL, VDCL, DCE, BNZ and CLFM (as well as for the compounds PCE and TCE) are given in Appendix 3. Application of the methods described above using these potency distributions yielded the station-specific distributions of increased risk given in Appendix 3, and the corresponding overall distribution of increased risk is also given in Appendix 3. From the latter distribution, we select the 50th percentile value equal to 1.5×10^{-6} , or approximately one chance in a million, as our best estimate of increased lifetime cancer risk for the average individual due to residence in the vicinity of the BKK landfill for the 7-year period of interest. Assuming a population of 7700 exposed individuals in the BKK landfill vicinity, this results in a corresponding predicted population risk of about 0.01 cases.

4. Discussion

Our calculation indicates that it is highly unlikely that even a single case of cancer would arise from the described BKK landfill exposure scenario. This result is consistent with the DHS report which suggested that 50×10^{-6} represents a "maximum risk" and that "the true excess risk of cancer is probably lower" (DHS 1983, p. 12).

Other various assumptions that could have led to a lower result were not included in our analysis. Such alternative assumptions include factors to take into account (1) seasonal adjustment of concentration data; (2) the possibility that the reported ambient concentration differences are due to chance such that the actual TWA concentrations at the landfill and at the comparison stations are equal and therefore the increased cancer risk is 0; (3) the possibility that the effective dilution factor due to atmospheric mixing might be much greater than the factor of 2 assumed for the average individual in the BKK residential population; or (4) the possibility that perhaps less than 50% of VDCL inhaled by humans will be metabolized even at low doses, in accordance with our analysis in Section 2.2.1.4 of data in EPA (1985b) relied on for VDCL potency assessment.

We did investigate, however, the effect of using alternative assumptions regarding (1) the carcinogenicity of PCE and TCE and (2) the relative population density in BKK-vicinity areas influenced by Stations A, B, D and F. The inclusion of PCE and TCE potencies into our risk analysis, corresponding with a presumption that these compounds are human carcinogens at low doses analogous to the other compounds VCL, VDCL, DCE, BENZ and CLFM, results in an increase in our "best" estimate of increased individual risk from 1.5×10^{-6} to 2.2×10^{-6} , that is, an increase by a factor of only 1.47. This result highlights the fact that increased risk in our analysis is primarily due to the impact of VDCL's carcinogenic potency, which was calculated to be greater than that of the other compounds. In the context of our calculation of overall increased risk using a weighted average of station-specific risks, the effect of changing the station weights from (.25,.25,.25,.25) to (.25,.40,.15,.20) for Stations A, B, D and F, respectively, was negligible.

To put these risks further into perspective it must be understood that a 1×10^{-6} lifetime mortality risk is roughly equivalent to 19 minutes of life expectancy, or put another way, equal to the added risk of death that is incurred by driving a car 100 miles, by riding a bicycle 10 miles, by traveling for 6 minutes by canoe or by smoking 1 or 2 cigarettes (cf. Crouch and Wilson 1952). The latter risk, smoking 1-2 cigarettes, is approximately equivalent in cancer risk to living for two weeks in Los Angeles as a result of air pollution (cf. Ames 1983). Residents living in the vicinity of the BKK landfill therefore experience only a negligible increased cancer risk, similar to that now being used by the U.S. Food and Drug Administration as a benchmark of insignificant health risk requiring no regulatory action (FDA 1985).

At very low doses, individual cancer risk becomes approximately equal to carcinogenic potency multiplied by dose, according to the cancer risk extrapolation model we have used in this analysis. Using the published EPA ("95% upper confidence limit") potency values, this relation allows us to make certain comparisons of estimated excess individual risk associated with different exposure situations. Note that such risk values are only estimates that contain a great deal of uncertainty, and only a *small part* of this uncertainty is meant to be addressed by EPA's use of "upper confidence limit" risk estimates; see, e.g., Anderson *et al.* (1983), Cothorn *et al.* (1984) and Bogen (1986). Nevertheless, such risk estimates are useful for comparative purposes, as we have undertaken here.

Firstly, we calculated the excess individual risk due to exposures (15 hours per day) at an average BKK residence (see Tables 2.1 and 2.3). We then compared this risk value of 2.4×10^{-6} with that associated with other typical exposures to some of the compounds listed in Table 2.1 and exposure to other "everyday" carcinogens such as aflatoxin B₁, which is present in all peanut

butter at an average of 2 ppb and which is regulated at 20 ppb (Table 2.4). One peanut butter sandwich per day works out to be at least 10 times worse than residence near BKK, which is similar in risk to drinking ordinary U.S. tap water and much less risky than taking a daily hot tub. This increased individual risk value of 2.4×10^{-6} is also absolutely minimal when compared to the risk incurred by someone working in a dry cleaning shop or in a plant manufacturing vinylidene chloride (Table 2.4). We therefore conclude that exposure of residents to the compounds and ambient concentrations listed in Table 2.1 poses a minimal, negligible cancer risk—a risk that is highly unlikely to result in even a single excess cancer case over the lifetime of the exposed BKK residential population.

REFERENCES

- Ames, B., 1983. "Dietary carcinogens and anticarcinogens." *Science*, vol. 221, pp. 1256-1264.
- Anderson, E. L., and the Carcinogen Assessment Group of the U.S. Environmental Protection Agency, 1983. "Quantitative approaches in use to assess cancer risk," *Risk Analysis*, vol. 3, no. 4, pp. 277-295.
- Bogen, K.T., 1986. *Uncertainty in Environmental Health Risk Assessment: A Framework for Analysis and an Application to a Chronic Exposure Situation Involving a Chemical Carcinogen*, Dissertation, School of Public Health, University of California at Berkeley.
- Cothorn, C.R., W.A. Coniglio, and W.L. Marcus, 1984. "Techniques for the Assessment of Carcinogenic Risk to the U.S. Population Due to Exposure from Selected Volatile Organic Compounds from Drinking Water Via the Ingestion, Inhalation and Dermal Routes," U.S. Environmental Protection Agency, Office of Drinking Water, EPA 570/9-85-001 (1 May 1984), NTIS Report No. PB84-213941, Washington D.C.
- Crouch, E.A.C. and R. Wilson, 1982, *Risk/Benefit Analysis*, Ballinger, Cambridge MA, pp. 165-193.
- Crump, K.S. and W.W. Watson, 1979, "GLOBAL79: A Fortran Program to Extrapolate Dichotomous Animal Carcinogenicity Data to Low Doses," National Institute of Environmental Health Sciences, Contract No. 1-ES-2123 (1979).
- Crump, K.S. and R.B. Howe, 1984, "The multistage model with a time-dependent dose pattern: applications to carcinogenic risk assessment," *Risk Analysis*, vol. 4, no. 3, pp. 163-176.
- DHS, California Department of Health Services, Toxic Substances Control Division; California Air Resources Board, Haagen-Smit Laboratory Division; and South Coast Air Quality Management District, 1983. "Ambient Air Monitoring and Health Risk Assessment for Suspect Human Carcinogens Around the BKK Landfill in West Covina," (March 1983).
- EPA, U.S. Environmental Protection Agency, 1984a, "Proposed Guidelines for Carcinogen Risk Assessment," *Federal Register*, vol. 49, no. 227 (23 Nov. 1984), pp. 46294-46301.
- EPA, U.S. Environmental Protection Agency, Office of Research and Development, Office of Health and Environmental Assessment, Environmental Criteria and Assessment Office, 1984b, "Health Assessment Document for Tetrachloroethylene (Perchloroethylene)," (EPA/600/8-82/005F Final Report, July 1985), Research Triangle Park, NC 27711.
- EPA, U.S. Environmental Protection Agency, Office of Research and Development, Office of Health and Environmental Assessment, Environmental Criteria and Assessment Office, 1985a, "Health Assessment Document for Trichloroethylene," (EPA/600/8-82/008F Final Report, July 1985; NTIS # PB85-249696), Research Triangle Park, NC 27711.
- EPA, U.S. Environmental Protection Agency, Office of Research and Development, Office of Health and Environmental Assessment, Environmental Criteria and Assessment Office, 1985b, "Health Assessment Document for Vinylidene Chloride," (EPA/600/8-83/031F, August 1985), Research Triangle Park, NC 27711.

- EPA, U.S. Environmental Protection Agency, Office of Research and Development, Office of Health and Environmental Assessment, Environmental Criteria and Assessment Office, 1985c, "Health Assessment Document for 1,2-Dichloroethane (Ethylene Chloride)," (EPA/600/8-84/006F Final Report, Sept. 1985), Research Triangle Park, NC 27711.
- EPA, U.S. Environmental Protection Agency, Office of Research and Development, Office of Health and Environmental Assessment, Environmental Criteria and Assessment Office, 1985d, "Health Assessment Document for Chloroform," (EPA/600/8-84/004F Final Report, Sept. 1985), Research Triangle Park, NC 27711.
- FDA, U.S. Food and Drug Administration, 1985, "Proposed ban on the use of methylene chloride as an ingredient of aerosol cosmetic products," *Federal Register*, vol. 50, no. (18 Dec. 1985), pp. 51551-51559.
- Fox, T.R. and P.G. Watanabe, 1985, "Detection of a cellular oncogene in spontaneous liver tumors of B6C3F1 mice," *Science*, vol. 229, pp. 596-597.
- Gehring, P.J., P.G. Watanabe, and C.N. Park, 1978, "Resolution of dose-response toxicity data for chemicals requiring metabolic activation: Example—Vinyl Chloride," *Toxicol. Appl. Pharmacol.*, vol. 44, pp. 581-591.
- Gehring, P.J., P.G. Watanabe, and C.N. Park, 1979, "Risk of angiosarcoma in workers exposed to vinyl chloride as predicted from studies in rats," *Toxicol. Appl. Pharmacol.*, vol. 49, pp. 15-21.
- Hogan, M.D. and D.G. Hoel, 1982, "Extrapolation to man," in *Principles and Methods of Toxicology*, ed. A.W. Hayes, Raven Press, New York, pp. 711-731.
- Maltoni, C. and G. LeFemine, 1975, "Carcinogenicity assays of vinyl chloride: current results," *Ann. N.Y. Acad. Sci.*, vol. 248, pp. 195-224.
- Maltoni, C., 1983, "Myths and facts in the history of benzene carcinogenicity," in *Advances in Environmental Toxicology, Vol. IV: Carcinogenicity and Toxicity of Benzene*, ed. M.A. Mehlman, Princeton Scientific Publishers Inc., Princeton, pp. 1-13.
- Mirsalis, J.C., C.K. Tyson, E.N. Loh, K.L. Steinmetz, J.P. Bakke, C.M. Hamilton, D.K. Spak, and J.W. Spalding, 1985, "Induction of hepatic cell proliferation and unscheduled DNA synthesis in mouse hepatocytes following in vivo treatment," *Carcinogenesis*, vol. 6, pp. 1521-1524.
- NAS, National Academy of Sciences, Committee on the Institutional Means for Assessment of Risks to Public Health, 1983, *Risk Assessment in the Federal Government: Managing the Process*, National Academy Press, Washington, D.C., pp. 36-37.
- NTP, National Toxicology Program, 1985, "NTP Technical Report on the Toxicology and Carcinogenesis Studies of Tetrachloroethylene (Perchloroethylene) (Cas No. 127-18-4) In F344/N Rats and B6C3F1 Mice (Inhalation Studies)," Draft Report dated 9/85 (NTP TR 311, NIH Publication No. 85-2567), Research Triangle Park, NC 27703.

Table 2.1

Increased Exposure Above Measured Control Levels for
Air Contaminants Measured at 4 BKK-Vicinity Monitoring Stations

Increased Exposure in mg/kg/day (assuming 20 m ³ /70 kg/day intake)					
At Monitoring Station:					
Compound	A	B	D	F	Mean
Vinyl Chloride	.0034	.0018	.0011	0.0	.0016
Perchloroethylene	.0019	.00087	.0015	.0011	.0013
Trichloroethylene	.0010	.0016	.0021	.0011	.0015
Vinylidene Chloride	.0011	.00074	.00062	.00017	.0007
1,2-Dichloroethane	.0019	.0014	.0014	.00023	.0012
Benzene	.0011	.00091	.00091	-.00055	.0006
Chloroform	0.0	.00007	.00056	-.00029	.00009

TABLE 2.2

POTENCY ANALYSIS INPUT

KEY: LINE #1 = TITLE
 LINE #2 = # ANIMALS PER DOSE GROUP (SURVIVING UNTIL
 AFTER THE OCCURRENCE OF THE 1ST TUMOR
 IN THE EXPERIMENT)
 LINE #3 = # TUMOR-BEARING ANIMALS PER DOSE GROUP
 LINE #4 = # EQUIVALENT APPLIED DOSES FOR EACH DOSE GROUP
 (IN MG/KG/DAY)

VCL MALTONI75 SPR-DAW RATS FEMALE .35KG HEAD MG/KG/D ZYMBAI GLAND
 58,59,59,59,59,60,61
 0,0,0,4,2,7,16
 0,0,4,225,13,92,19,52,28,76,30,90,31,57

VCL MALTONI75 SPR-DAW RATS FEMALE .35KG HEAD MG/KG/D NEPHROELASTOMAS
 58,59,59,59,59,60,61
 0,1,0,4,6,4,5
 0,0,4,225,13,92,19,52,28,76,30,90,31,57

VCL MALTONI75 SPR-DAW RATS FEMALE .35KG HEAD PG/KG/D LIVER ANGIOSARCOMAS
 58,59,59,59,59,60,61
 0,1,4,7,3,13,9
 0,0,4,225,13,92,19,52,28,76,30,90,31,57

CL MALTONI75 SPR-DAW RATS FEMALE .35KG HEAD MG/KG/D COMPOSITE
 58,59,59,59,59,60,61
 6,10,10,22,32,31,38
 0,0,4,225,13,92,19,52,28,76,30,90,31,57

BENZENE MALTONI83 S-C RATS FEMALE .30KG HEAD MG/KG/D ZYMBAI GLAND
 30,30,32
 0,2,8
 0,0,16,07,80,36

VINYLDENE CL MALTONI85 SW MICE .035KG HEAD MG/KG/D KIDNEY ADENOCARC.
 126,25,119
 0,0,28
 0,0,0,9627,2,457

DCE NC178 D-M RATS MALE .5KG HEAD MG/KG/D HEPANGIOSARCOMAS
 40,46,27
 0,9,7
 0,0,23,10,42,75

DCE NC178 B6C3F1 MICE MALE .035KG HEAD MG/KG/D HEPATOCARCINOMAS
 19,47,46
 1,6,12
 0,0,44,35,57,70

HLGRDFORM NC176 B6C3F1 MICE MALE.035KG HEAD MG/KG/D HEPATCCARCINOMAS
 50.45
 1.8.44
 0.0.79.42,159.4

CHLGRDFORM NC176 B6C3F1 MICE FEMALE.028KG HEAD MG/KG/D HEPATCCARCINOMAS
 20.45.41
 0.36.39
 0.0.137.0.274.5

PCE NC177 B6C3F1 MICE, MALES HALD (MG/KG/DAY) VIA SAME LIVER
 20.48.45
 2.32.27
 0.0.118.66.163.89

PCE NC177 B6C3F1 MICE, FEMALES HALD (MG/KG/DAY) VIA SAME LIVER
 20.48.45
 0.19.19
 0.0.97.718.142.76

PCE NTP85 B6C3F1 MICE: MALE HALD MG/KG/DAY VIA SAME,HEP AD/CARC
 47.46.50
 16.31.40
 0.0.117.96.203.02

PCE NTP85 B6C3F1 MICE: FEMALE HALD MG/KG/DAY VIA SAME, HEP CARC
 9.40.46
 13.26
 .124.79.209.89

TCE NTP82 B6C3F1 MICE: MALE HALD MG/KG/DAY VIA SAME HEP CARC
 48.50
 8.30
 0.0.675.830

TCE NTP82 B6C3F1 MICE: FEMALE HALD MG/KG/DAY VIA SAME HEP CARC
 48.49
 2.13
 0.0.680.409

TCE NC176 B6C3F1 MICE: MALES HALD MG/KG/DAY VIA SAME HEP CARC
 20.50.48
 1.26.31
 0.0.444.595.845.545

TCE NC176 B6C3F1 MICE: FEMALES HALD MG/KG/DAY VIA SAME HEP CARC
 20.50.47
 0.4.11
 0.0.337.779.655.510

APPENDIX THREE

RISK ANALYSIS OUTPUT

COL MALTION 75 S-D RATS FEMALE .35RG MG/MO'D ZYMBAI GLAND

Tumorigenic Potency Distribution Calculated
Assuming Equipotent Doses Can Be Extrapolated
Between Species on a (1) mg per Body Weight (BW)
Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/[mg/kg/day]) Calculated Using Extrapolation Assumption:		Percentile
	BW	SA	
0.4300	0.	0.	0.4300
1.0000	0.0005726795	0.0033490527	1.0000
2.0000	0.0012998717	0.0076017021	2.0000
3.0000	0.0018156692	0.0106181065	3.0000
4.0000	0.0021266260	0.0124365935	4.0000
5.0000	0.0023978178	0.0140225347	5.0000
6.0000	0.0026284026	0.0153710032	6.0000
7.0000	0.0028122789	0.0164463203	7.0000
8.0000	0.0029672480	0.0173525847	8.0000
9.0000	0.0031151413	0.0182174705	9.0000
10.0000	0.0032536136	0.0190272629	10.0000
11.0000	0.0033952545	0.0198555849	11.0000
12.0000	0.0035280201	0.0206320025	12.0000
13.0000	0.0036386538	0.0212789942	13.0000
14.0000	0.0037545366	0.0219566803	14.0000
15.0000	0.0038785900	0.0226621490	15.0000
16.0000	0.0039749783	0.0232458320	16.0000
17.0000	0.0040668273	0.0237946641	17.0000
18.0000	0.0041523678	0.0242832135	18.0000
19.0000	0.0042639095	0.0249355137	19.0000
20.0000	0.0043675330	0.0255418010	20.0000
21.0000	0.0044624414	0.0260965358	21.0000
22.0000	0.0045404690	0.0265528448	22.0000
23.0000	0.0046295142	0.0270735845	23.0000
24.0000	0.0047117588	0.0275545549	24.0000
25.0000	0.0047984691	0.0280616395	25.0000
26.0000	0.0048768963	0.0285202842	26.0000
27.0000	0.0049669673	0.0290470235	27.0000
28.0000	0.0050474182	0.0295175035	28.0000
29.0000	0.0051244469	0.0299679711	29.0000
30.0000	0.0051969453	0.0303919446	30.0000
31.0000	0.0052697374	0.0308176354	31.0000
32.0000	0.0053448598	0.0312569551	32.0000
33.0000	0.0054183519	0.0316867361	33.0000
34.0000	0.0054830536	0.0320651159	34.0000
35.0000	0.0055457177	0.0324315801	35.0000
36.0000	0.0056098541	0.0328066535	36.0000
37.0000	0.0056853541	0.0332481787	37.0000
38.0000	0.0057550259	0.0336556211	38.0000
39.0000	0.0058230604	0.0340534896	39.0000
40.0000	0.0058844727	0.0344412636	40.0000
41.0000	0.0059575361	0.0348402150	41.0000
42.0000	0.0060177930	0.0351922959	42.0000
43.0000	0.0060733937	0.0355203152	43.0000
44.0000	0.0061279945	0.0358367562	44.0000
45.0000	0.0061864941	0.0361768645	45.0000
46.0000	0.0062544325	0.0365764648	46.0000
47.0000	0.0063161105	0.0369368680	47.0000
48.0000	0.0063730385	0.0372697860	48.0000
49.0000	0.0064345961	0.0376297757	49.0000
50.0000	0.0064955465	0.0379862152	50.0000
51.0000	0.0065528918	0.0383215733	51.0000
52.0000	0.0066151384	0.0386468821	52.0000

55.0000	0.0067972252	0.0397504456	55.0000
56.0000	0.0068522769	0.0400758962	56.0000
57.0000	0.0069182469	0.0404581837	57.0000
58.0000	0.0069725960	0.0407760218	58.0000
59.0000	0.0070274523	0.0410966214	59.0000
60.0000	0.0070795454	0.0414014645	60.0000
61.0000	0.0071384124	0.0417457223	61.0000
62.0000	0.0071989144	0.0420995392	62.0000
63.0000	0.0072529400	0.0424154848	63.0000
64.0000	0.0073079383	0.0427371152	64.0000
65.0000	0.0073671667	0.0430834852	65.0000
66.0000	0.0074252975	0.0434234366	66.0000
67.0000	0.0074962992	0.0438386574	67.0000
68.0000	0.0075497241	0.0441510901	68.0000
69.0000	0.0076115336	0.0445125550	69.0000
70.0000	0.0076629536	0.0448132604	70.0000
71.0000	0.0077161331	0.0451362431	71.0000
72.0000	0.0077770394	0.0454604376	72.0000
73.0000	0.0078272959	0.0457743406	73.0000
74.0000	0.0078862319	0.0461190008	74.0000
75.0000	0.0079470577	0.04646747697	75.0000
76.0000	0.0080047920	0.0468123443	76.0000
77.0000	0.0080706535	0.0471975654	77.0000
78.0000	0.0081326142	0.0475598536	78.0000
79.0000	0.0082012340	0.0479611456	79.0000
80.0000	0.0082558179	0.0482803546	80.0000
81.0000	0.0083276853	0.0487006381	81.0000
82.0000	0.0083982218	0.0491131395	82.0000
83.0000	0.0084746592	0.0495601483	83.0000
84.0000	0.0085414629	0.0499508195	84.0000
85.0000	0.0086164356	0.0503895544	85.0000
86.0000	0.0086853262	0.0507950708	86.0000
87.0000	0.0087648258	0.0512570515	87.0000
88.0000	0.0088466890	0.0517357523	88.0000
89.0000	0.0089489324	0.0522337163	89.0000
90.0000	0.0090379392	0.0528536476	90.0000
91.0000	0.0091302823	0.0533942580	91.0000
92.0000	0.0092368033	0.0540171973	92.0000
93.0000	0.0093518551	0.0546900257	93.0000
94.0000	0.0094750915	0.0554107167	94.0000
95.0000	0.0096185394	0.0562496036	95.0000
96.0000	0.0097883055	0.0572424047	96.0000
97.0000	0.0099880612	0.0584105812	97.0000
98.0000	0.0102364104	0.0598629378	98.0000
99.0000	0.0104925238	0.0625303090	99.0000

PCE NTP85 6003F1 MICE: FEMALES MALE (MG/KG/DAY) VIA SAME LIVER
 PCE NTP85 66C3 MICE: MALE MALE MG/KG/DAY IA SAME, HEP AD/CARC
 PCE NTP85 6003F1 MICE: FEMALE MALE MG/KG/DAY VIA SAME, HEP CARC

Tumorigenic Potency Distribution Calculated
 Assuming Equipotent Doses Can Be Extrapolated
 Between Species on a (1) mg per Body Weight (BW)
 Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/[mg/kg/day]) Calculated Using Extrapolation Assumption:		Percentile
	BW	SA	
23.5200	0.	0.	23.5200
24.0000	0.0000631399	0.0008419139	24.0000
25.0000	0.0002477595	0.0032905159	25.0000
26.0000	0.0004065095	0.0053327973	26.0000
27.0000	0.0005406287	0.0071042348	27.0000
28.0000	0.0006775936	0.0089232866	28.0000
29.0000	0.0008084442	0.0105216913	29.0000
30.0000	0.0009508039	0.0124029778	30.0000
31.0000	0.0010622796	0.0140911490	31.0000
32.0000	0.0012121021	0.0157920271	32.0000
33.0000	0.0013783769	0.0162200354	33.0000
34.0000	0.0015607907	0.0207534954	34.0000
35.0000	0.0017147622	0.0223614757	35.0000
36.0000	0.0018365651	0.0239088777	36.0000
37.0000	0.0019815555	0.0262096580	37.0000
38.0000	0.0021057397	0.0275176186	38.0000
39.0000	0.0022612270	0.0296708245	39.0000
40.0000	0.0023447108	0.0310381241	40.0000
41.0000	0.0024843295	0.0326073261	41.0000
42.0000	0.0026001076	0.0346589722	42.0000
43.0000	0.0026962133	0.0360331982	43.0000
44.0000	0.0028273263	0.0368654616	44.0000
45.0000	0.0029609974	0.0390280671	45.0000
46.0000	0.0030838503	0.0408094414	46.0000
47.0000	0.0032178843	0.0419077836	47.0000
48.0000	0.0033038284	0.0438546464	48.0000
49.0000	0.0033742981	0.0451851636	49.0000
50.0000	0.0034405054	0.0462058075	50.0000
51.0000	0.0035401534	0.0471385531	51.0000
52.0000	0.0036381371	0.0481932069	52.0000
53.0000	0.0036876083	0.0497919470	53.0000
54.0000	0.0037836900	0.0509260669	54.0000
55.0000	0.00386175124	0.0517350100	55.0000
56.0000	0.0039036532	0.0527276520	56.0000
57.0000	0.0039779469	0.0537138171	57.0000
58.0000	0.0040559196	0.0543200411	58.0000
59.0000	0.0041189343	0.0555971563	59.0000
60.0000	0.0041798120	0.0560675710	60.0000
61.0000	0.0042497516	0.0569670014	61.0000
62.0000	0.0043027191	0.0578635931	62.0000
63.0000	0.0043729548	0.0585039258	63.0000
64.0000	0.0044210218	0.0596585087	64.0000
65.0000	0.0044639323	0.0602021478	65.0000
66.0000	0.0045630080	0.0613929220	66.0000
67.0000	0.0046134321	0.0620164797	67.0000
68.0000	0.0047130641	0.0625100061	68.0000
69.0000	0.0047519526	0.0634471178	69.0000
70.0000	0.0048404224	0.0644137169	70.0000
71.0000	0.0049026082	0.0649361536	71.0000
72.0000	0.0049709054	0.0663178042	72.0000
73.0000	0.0050370945	0.0668002702	73.0000

76.0000	0.0052273550	0.0689715445	76.0000
77.0000	0.00531083	0.0696585774	77.0000
78.0000	0.0053876548	0.0707354471	78.0000
79.0000	0.0054476908	0.0713223070	79.0000
80.0000	0.0055350373	0.0723741651	80.0000
81.0000	0.0056042060	0.0732360450	81.0000
82.0000	0.0056612912	0.0739637436	82.0000
83.0000	0.0057923449	0.0750207230	83.0000
84.0000	0.0058550616	0.0761620775	84.0000
85.0000	0.0059278943	0.0772702247	85.0000
86.0000	0.0060320524	0.0783423260	86.0000
87.0000	0.0061395695	0.0797456056	87.0000
88.0000	0.0062557189	0.0812087208	88.0000
89.0000	0.0063808989	0.0824906453	89.0000
90.0000	0.0064814021	0.0840217695	90.0000
91.0000	0.0066168713	0.0853406787	91.0000
92.0000	0.0067779957	0.0867500360	92.0000
93.0000	0.0068894438	0.0882896659	93.0000
94.0000	0.0070212791	0.0911299139	94.0000
95.0000	0.0072206082	0.0928635597	95.0000
96.0000	0.0073681385	0.0958132222	96.0000
97.0000	0.0075680926	0.0984796148	97.0000
98.0000	0.0078990255	0.1026449278	98.0000
99.0000	0.0084332814	0.1077851206	99.0000

TCE NCI76 B6C3F1 MICE: MALES HALD MG/KG/DAY VIA SAME HEP CARC
TCE NCI76 B6C3F1 MICE: FEMALES HALD MG/KG/DAY VIA SAME HEP CARC

Tumorigenic Potency Distribution Calculated
Assuming Equipotent Doses Can Be Extrapolated
Between Species on a (1) mg per Body Weight (BW)
Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/[mg/kg/day]) Calculated Using Extrapolation Assumption:		Percentile
	BW	SA	
11.6000	0.	0.	11.6000
12.0000	0.0000298718	0.0004010E05	12.0000
13.0000	0.0000331952	0.0004617967	13.0000
14.0000	0.0000735215	0.0010205493	14.0000
15.0000	0.0000768560	0.0010693242	15.0000
16.0000	0.0001185921	0.0016247210	16.0000
17.0000	0.0001299646	0.0016874E71	17.0000
18.0000	0.0001623041	0.0021598E54	18.0000
19.0000	0.0001717535	0.0022996233	19.0000
20.0000	0.0001994221	0.0025248E00	20.0000
21.0000	0.0002066782	0.0027376411	21.0000
22.0000	0.0002111395	0.0028E02075	22.0000
23.0000	0.0002317257	0.0029799310	23.0000
24.0000	0.0002410785	0.0030773962	24.0000
25.0000	0.0002555447	0.0033E21E20	25.0000
26.0000	0.0002674929	0.0034916122	26.0000
27.0000	0.0002729225	0.0035558483	27.0000
28.0000	0.0002864209	0.0037451424	28.0000
29.0000	0.0002987948	0.0038907297	29.0000
30.0000	0.0003045304	0.0040064328	30.0000
31.0000	0.0003124588	0.0040976494	31.0000
32.0000	0.0003182912	0.00417656E7	32.0000
33.0000	0.0003278030	0.0043178336	33.0000
34.0000	0.0003427067	0.00441341E5	34.0000
35.0000	0.0003502932	0.0044350E01	35.0000
36.0000	0.0003569786	0.0046192929	36.0000
37.0000	0.0003671742	0.0048116501	37.0000
38.0000	0.0003819009	0.0048415521	38.0000
39.0000	0.0003870564	0.0049207699	39.0000
40.0000	0.0003929799	0.0050728549	40.0000
41.0000	0.0004038969	0.0052447678	41.0000
42.0000	0.0004203070	0.0054401E90	42.0000
43.0000	0.0004319644	0.0055211429	43.0000
44.0000	0.0004422651	0.0057088E00	44.0000
45.0000	0.0004531117	0.0058406456	45.0000
46.0000	0.0004652110	0.0059791761	46.0000
47.0000	0.0004817551	0.0061627184	47.0000
48.0000	0.0005044929	0.0063774111	48.0000
49.0000	0.0005184426	0.006542694E	49.0000
50.0000	0.0005371177	0.0067672594	50.0000
51.0000	0.0005564941	0.0070168427	51.0000
52.0000	0.0005792555	0.0073040705	52.0000
53.0000	0.0006026781	0.0075932685	53.0000
54.0000	0.0006419369	0.0080280732	54.0000
55.0000	0.0006660301	0.0083375992	55.0000
56.0000	0.0006980940	0.0086673E99	56.0000
57.0000	0.0007303558	0.0089960C33	57.0000
58.0000	0.0007553491	0.0093287C15	58.0000
59.0000	0.0007789282	0.009641750E	59.0000
60.0000	0.0008099292	0.0099793402	60.0000
61.0000	0.0008371107	0.0103200000	61.0000

62.0000	0.C008 4004	0.01059529e5	62.00
63.0000	0.0006792258	0.0109503596	63.0000
64.0000	0.0009100983	0.0111167086	64.0000
65.0000	0.0009260917	0.0113879135	65.0000
66.0000	0.0009449993	0.0116341598	66.0000
67.0000	0.0009639515	0.0118264370	67.0000
68.0000	0.0009767214	0.0120875e96	68.0000
69.0000	0.0009926786	0.0122434404	69.0000
70.0000	0.0010138333	0.0126162264	70.0000
71.0000	0.0010469393	0.0126998471	71.0000
72.0000	0.0010547247	0.0130873E81	72.0000
73.0000	0.0010860262	0.0132064130	73.0000
74.0000	0.0010860262	0.0135276621	74.0000
75.0000	0.0011097921	0.0136578409	75.0000
76.0000	0.0011233548	0.0139594963	76.0000
77.0000	0.0011408336	0.0140514e16	77.0000
78.0000	0.0011601131	0.0143212499	78.0000
79.0000	0.0011717480	0.0144456E86	79.0000
80.0000	0.0011930985	0.0146767264	80.0000
81.0000	0.0012043920	0.0149525125	81.0000
82.0000	0.0012199676	0.0152541241	82.0000
83.0000	0.0012353265	0.0155025935	83.0000
84.0000	0.0012502574	0.0156682275	84.0000
85.0000	0.0012755778	0.0159120e34	85.0000
86.0000	0.0012912479	0.0160E74929	86.0000
87.0000	0.0013026590	0.0163902510	87.0000
88.0000	0.0013261532	0.01661e5269	88.0000
89.0000	0.001338998E	0.0167716797	89.0000
90.0000	0.0013619312	0.0170479119	90.0000
91.0000	0.0013769770	0.0172e18143	91.0000
92.0000	0.0013963441	0.0176176969	92.0000
93.0000	0.0014159299	0.0179115e27	93.0000
94.0000	0.0014415335	0.0181E49767	94.0000
95.0000	0.0014810515	0.0185e76906	95.0000
96.0000	0.0015181559	0.0190e31E78	96.0000
97.0000	0.0015631353	0.0197C35E75	97.0000
98.0000	0.0016321145	0.0204428229	98.0000
99.0000	0.00175341e7	0.0217e68433	99.0000

VINYLDENE CL MALTON 5 SW MICE .035KG MG/KG/D KID Y ADENOCARCINOMAS

Tumorigenic Potency Distribution Calculated
Assuming Equipotent Doses Can Be Extrapolated
Between Species on a (1) mg per Body Weight (BW)
Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/[mg/kg/day]) Calculated Using Extrapolation Assumptions:		Percentile
	BW	SA	
57.4000	0.	0.	57.4000
58.0000	0.0023508139	0.0296163750	58.0000
59.0000	0.0023508139	0.0296163750	59.0000
60.0000	0.0023508139	0.0296163750	60.0000
61.0000	0.0023508139	0.0296163750	61.0000
62.0000	0.0052523129	0.0661749393	62.0000
63.0000	0.0052523129	0.0661749393	63.0000
64.0000	0.0052523129	0.0661749393	64.0000
65.0000	0.0081231073	0.1023446545	65.0000
66.0000	0.0081231073	0.1023446545	66.0000
67.0000	0.0109638385	0.1381355971	67.0000
68.0000	0.0109638385	0.1381355971	68.0000
69.0000	0.0137751363	0.1735557020	69.0000
70.0000	0.0137751363	0.1735557020	70.0000
71.0000	0.0165575966	0.2086124718	71.0000
72.0000	0.0193118080	0.2433133423	72.0000
73.0000	0.0247377418	0.3116757572	73.0000
74.0000	0.0403775983	0.5087254643	74.0000
75.0000	0.0501255058	0.6315413117	75.0000
76.0000	0.0532964957	0.6715164450	76.0000
77.0000	0.0595352550	0.7500967383	77.0000
78.0000	0.0626006722	0.7887184024	78.0000
79.0000	0.0626006722	0.7887184024	79.0000
80.0000	0.0656318665	0.8269090056	80.0000
81.0000	0.0656318665	0.8269090056	81.0000
82.0000	0.0686295033	0.8646768332	82.0000
83.0000	0.0715944767	0.9020331502	83.0000
84.0000	0.0715944767	0.9020331502	84.0000
85.0000	0.0745273232	0.9389846921	85.0000
86.0000	0.0745273232	0.9389846921	86.0000
87.0000	0.0774288177	0.9755411744	87.0000
88.0000	0.0774288177	0.9755411744	88.0000
89.0000	0.0794975162	1.0016051531	89.0000
90.0000	0.0802996159	1.0117108822	90.0000
91.0000	0.0831080675	1.0470951796	91.0000
92.0000	0.0833316445	1.0499120951	92.0000
93.0000	0.0833316445	1.0499120951	93.0000
94.0000	0.0908618569	1.1447867155	94.0000
95.0000	0.0949393337	1.2409447432	95.0000
96.0000	0.1068254709	1.3459155560	96.0000
97.0000	0.1109157301	1.3974500695	97.0000
98.0000	0.1188809972	1.4978053570	98.0000
99.0000	0.1234383583	1.5552245378	99.0000

Tumorigenic Potency Distribution Calculated
Assuming Equipotent Doses Can be Extrapolated
Between Species on a (1) mg per Body Weight (BW)
Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/(mg/kg/day)) Calculated Using Extrapolation Assumption:		Percentile
	BW	SA	
32.0550	0.	0.	32.0550
33.0000	0.0001038053	0.0010927507	33.0000
34.0000	0.0004158972	0.0034300131	34.0000
35.0000	0.0006223086	0.0052099419	35.0000
36.0000	0.0008052544	0.0067426893	36.0000
37.0000	0.0010680376	0.0078208055	37.0000
38.0000	0.0012650864	0.0101002697	38.0000
39.0000	0.0013524315	0.0106902849	39.0000
40.0000	0.0014808926	0.0128090139	40.0000
41.0000	0.0016244322	0.0142214093	41.0000
42.0000	0.0016513089	0.0148719465	42.0000
43.0000	0.0020176996	0.0160176922	43.0000
44.0000	0.0020587973	0.0172541905	44.0000
45.0000	0.0022687374	0.0193555558	45.0000
46.0000	0.0023955372	0.0192078575	46.0000
47.0000	0.0025417092	0.0205113776	47.0000
48.0000	0.0026377053	0.0220622046	48.0000
49.0000	0.0027363470	0.0222374797	49.0000
50.0000	0.0028409075	0.0227661040	50.0000
51.0000	0.0028641261	0.0244304445	51.0000
52.0000	0.0030001164	0.0251966689	52.0000
53.0000	0.0031688850	0.0264859651	53.0000
54.0000	0.0033263485	0.0266618678	54.0000
55.0000	0.0034470086	0.0272008087	55.0000
56.0000	0.0035790491	0.0280365832	56.0000
57.0000	0.0036291692	0.0291268602	57.0000
58.0000	0.0037549126	0.0296260266	58.0000
59.0000	0.0039007415	0.0300076716	59.0000
60.0000	0.0041051777	0.0307655465	60.0000
61.0000	0.0041463746	0.0314085409	61.0000
62.0000	0.0042734761	0.0326598287	62.0000
63.0000	0.0042826235	0.0329465419	63.0000
64.0000	0.0043652970	0.0335101448	64.0000
65.0000	0.0045969049	0.0340721458	65.0000
66.0000	0.0047065211	0.0351662982	66.0000
67.0000	0.0046733628	0.0355613206	67.0000
68.0000	0.0050965096	0.0356993228	68.0000
69.0000	0.0051347017	0.0363942012	69.0000
70.0000	0.0052294573	0.0382859968	70.0000
71.0000	0.0052772732	0.0386370271	71.0000
72.0000	0.0056311530	0.0387365334	72.0000
73.0000	0.0057518049	0.0390550457	73.0000
74.0000	0.0057926076	0.0395760536	74.0000
75.0000	0.0059661344	0.0404470563	75.0000
76.0000	0.0063385554	0.0412730231	76.0000
77.0000	0.0064640131	0.0415324606	77.0000
78.0000	0.0067303958	0.0426625122	78.0000
79.0000	0.0068540093	0.0434967284	79.0000
80.0000	0.0068751834	0.0443399596	80.0000
81.0000	0.0073733404	0.0445330563	81.0000
82.0000	0.0074477480	0.0450259484	82.0000
83.0000	0.0074609537	0.0455429740	83.0000

85.0000	C1 79178447	0.0476443917	85.0000
86.0000	0.0079985533	0.0484076105	86.0000
87.0000	0.0081310456	0.0498319604	87.0000
88.0000	0.0084179109	0.0504848696	88.0000
89.0000	0.0085530340	0.0508543812	89.0000
90.0000	0.0085997019	0.0517573170	90.0000
91.0000	0.0090888273	0.0522902869	91.0000
92.0000	0.0091861417	0.0535886660	92.0000
93.0000	0.0096188542	0.0543918086	93.0000
94.0000	0.0097569674	0.0557134598	94.0000
95.0000	0.0098993011	0.0570824519	95.0000
96.0000	0.0103204176	0.0582462363	96.0000
97.0000	0.0105662909	0.0606013367	97.0000
98.0000	0.0109909372	0.0631922334	98.0000
99.0000	0.0117450356	0.0666660815	99.0000

Tumorigenic Potency Distribution Calculated
Assuming Equipotent Doses Can Be Extrapolated
Between Species on a (1) mg per Body Weight (BW)
Basis or (2) mg per Surface Area (SA) Basis

Percentile	Potency (1/[mg/kg/day]) Calculated Using Extrapolation Assumptions:		Percentile
	BW	SA	
17.6000	0.	0.	17.3000
18.0000	0.0005611179	0.0035775830	18.0000
19.0000	0.0010148041	0.0062475200	19.0000
20.0000	0.0010154368	0.0062514152	20.0000
21.0000	0.0011749854	0.0072336569	21.0000
22.0000	0.0013267462	0.0081679663	22.0000
23.0000	0.0013267462	0.0081679663	23.0000
24.0000	0.0013685476	0.0084252951	24.0000
25.0000	0.0014714494	0.0090588015	25.0000
26.0000	0.0014714494	0.0090588015	26.0000
27.0000	0.0014714494	0.0090588015	27.0000
28.0000	0.0014714494	0.0090588015	28.0000
29.0000	0.0016097173	0.0099100312	29.0000
30.0000	0.0016097173	0.0099100312	30.0000
31.0000	0.0016097173	0.0099100312	31.0000
32.0000	0.0016097173	0.0099100312	32.0000
33.0000	0.0016097173	0.0099100312	33.0000
34.0000	0.0017354523	0.0106841046	34.0000
35.0000	0.0017420996	0.0107250279	35.0000
36.0000	0.0017420996	0.0107250279	36.0000
37.0000	0.0017420996	0.0107250279	37.0000
38.0000	0.0017420996	0.0107250279	38.0000
39.0000	0.0017420996	0.0107250279	39.0000
40.0000	0.0017420996	0.0107250279	40.0000
41.0000	0.0018690773	0.0115067502	41.0000
42.0000	0.0018690773	0.0115067502	42.0000
43.0000	0.0018690773	0.0115067502	43.0000
44.0000	0.0018690773	0.0115067502	44.0000
45.0000	0.0018690773	0.0115067502	45.0000
46.0000	0.0019910738	0.0122578079	46.0000
47.0000	0.0019910738	0.0122578079	47.0000
48.0000	0.0019910738	0.0122578079	48.0000
49.0000	0.0020652630	0.0127145443	49.0000
50.0000	0.0021053532	0.0129614165	50.0000
51.0000	0.0021084659	0.0129805179	51.0000
52.0000	0.0021084659	0.0129805179	52.0000
53.0000	0.0024735923	0.0152283749	53.0000
54.0000	0.0024735923	0.0152283749	54.0000
55.0000	0.0024735923	0.0152283749	55.0000
56.0000	0.0025360672	0.0174599085	56.0000
57.0000	0.0028703806	0.0176711548	57.0000
58.0000	0.0028703806	0.0176711548	58.0000
59.0000	0.0028703806	0.0176711548	59.0000
60.0000	0.0028703806	0.0176711548	60.0000
61.0000	0.0032219272	0.0198354088	61.0000
62.0000	0.0032400019	0.0199466632	62.0000
63.0000	0.0032801141	0.0201936290	63.0000
64.0000	0.0032801141	0.0201936290	64.0000
65.0000	0.0032801141	0.0201936290	65.0000
66.0000	0.0034188544	0.0210477673	66.0000
67.0000	0.0036572944	0.0225156955	67.0000
68.0000	0.0036572944	0.0225156955	68.0000
69.0000	0.0036572944	0.0225156955	69.0000
70.0000	0.0036572944	0.0225156955	70.0000

71.0000	0.00370 0034	0.0228011589	71.0000
72.0000	0.0037036634	0.0228011589	72.0000
73.0000	0.0037036634	0.0228011589	73.0000
74.0000	0.0037036634	0.0228011589	74.0000
75.0000	0.0037449196	0.0230551492	75.0000
76.0000	0.0037449196	0.0230551492	76.0000
77.0000	0.0040362515	0.0248486977	77.0000
78.0000	0.0040562302	0.0249716956	78.0000
79.0000	0.0040562302	0.0249716956	79.0000
80.0000	0.0040688526	0.0251725316	80.0000
81.0000	0.0040688526	0.0251725316	81.0000
82.0000	0.0040688526	0.0251725316	82.0000
83.0000	0.0041419901	0.0254996661	83.0000
84.0000	0.0041419901	0.0254996661	84.0000
85.0000	0.0041419901	0.0254996661	85.0000
86.0000	0.0042009316	0.0258625317	86.0000
87.0000	0.0042009316	0.0258625317	87.0000
88.0000	0.0042009316	0.0258625317	88.0000
89.0000	0.0044769049	0.0275615286	89.0000
90.0000	0.0045356899	0.0279234312	90.0000
91.0000	0.0045356899	0.0279234312	91.0000
92.0000	0.0049324445	0.0303660035	92.0000
93.0000	0.0049989149	0.0307752211	93.0000
94.0000	0.0049989149	0.0307752211	94.0000
95.0000	0.0054797754	0.0337355807	95.0000
96.0000	0.0054797754	0.0337355807	96.0000
97.0000	0.0056957078	0.0362962186	97.0000
98.0000	0.0059796646	0.0368130678	98.0000
99.0000	0.0064061992	0.0394389565	99.0000

80.0000	0.0107100270	0.1454702020	80.0000
81.0000	0.0108 3984	0.1472061574	81.0000
82.0000	0.0109540522	0.1486694813	82.0000
83.0000	0.0110216215	0.1493060440	83.0000
84.0000	0.0111465789	0.1512824744	84.0000
85.0000	0.0111928107	0.1512824744	85.0000
86.0000	0.0114462450	0.1553495824	86.0000
87.0000	0.0114462450	0.1553495824	87.0000
88.0000	0.0116167953	0.1573326141	88.0000
89.0000	0.0116897561	0.1586545406	89.0000
90.0000	0.0117840376	0.1595713543	90.0000
91.0000	0.0119430088	0.1620917022	91.0000
92.0000	0.0121220453	0.1640242189	92.0000
93.0000	0.0124753751	0.1689127068	93.0000
94.0000	0.0124858618	0.1694593579	94.0000
95.0000	0.0130839720	0.1775769740	95.0000
96.0000	0.0134910503	0.1830649078	96.0000
97.0000	0.0136579064	0.1858780384	97.0000
98.0000	0.0140798213	0.1910927445	98.0000
99.0000	0.0144791398	0.1965127140	99.0000

COMPOUNDS: VCL, VCCL, DCE, ENZ, CLFM

Calculated Distributions of Individual Risk

Percentile	STATION A	STATION B	STATION C	STATION D
0.0000	0.	0.	0.	-0.000060
1.0000	0.0000117384	0.0000074320	0.0000068312	-0.000057
2.0000	0.0000149576	0.0000093885	0.0000084504	-0.000054
3.0000	0.0000171173	0.0000105254	0.0000094657	-0.000051
4.0000	0.0000166231	0.0000114856	0.0000101038	-0.000049
5.0000	0.0000202142	0.0000123879	0.0000108344	-0.000046
6.0000	0.0000221594	0.0000130836	0.0000115001	-0.000045
7.0000	0.0000228705	0.0000137458	0.0000121013	-0.000043
8.0000	0.0000238056	0.0000143756	0.0000126325	-0.000041
9.0000	0.0000247567	0.0000149180	0.0000131480	-0.000039
10.0000	0.0000257650	0.0000155280	0.0000136969	-0.000038
11.0000	0.0000265681	0.0000159862	0.0000142634	-0.000035
12.0000	0.0000274636	0.0000165288	0.0000148324	-0.000034
13.0000	0.0000282500	0.0000169985	0.0000153391	-0.000032
14.0000	0.0000290385	0.0000175196	0.0000158206	-0.000030
15.0000	0.0000297792	0.0000180047	0.0000162584	-0.000028
16.0000	0.0000304257	0.0000183946	0.0000166756	-0.000025
17.0000	0.0000310246	0.0000188916	0.0000171831	-0.000023
18.0000	0.0000318354	0.0000193605	0.0000176232	-0.000022
19.0000	0.0000325451	0.0000198277	0.0000180155	-0.000019
20.0000	0.0000332134	0.0000203269	0.0000184686	-0.000016
21.0000	0.0000338986	0.0000208272	0.0000189382	-0.000014
22.0000	0.0000346417	0.0000213097	0.0000194253	-0.000012
23.0000	0.0000354209	0.0000218615	0.0000199231	-0.000011
24.0000	0.0000363165	0.0000224705	0.0000205214	-0.000009
25.0000	0.0000371377	0.0000228546	0.0000210565	-0.000007
26.0000	0.0000381047	0.0000235390	0.0000217582	-0.000006
27.0000	0.0000390199	0.0000242334	0.0000223537	-0.000005
28.0000	0.0000400237	0.0000250638	0.0000230269	-0.000005
29.0000	0.0000412684	0.0000258034	0.0000236835	-0.000004
30.0000	0.0000423530	0.0000264956	0.0000244092	-0.000004
31.0000	0.0000437418	0.0000274218	0.0000252453	-0.000004
32.0000	0.0000451495	0.0000284245	0.0000261823	-0.000004
33.0000	0.0000470129	0.0000294835	0.0000271678	-0.000004
34.0000	0.0000486159	0.0000305580	0.0000283033	-0.000004
35.0000	0.0000510023	0.0000324262	0.0000297674	-0.000004
36.0000	0.0000544181	0.0000349360	0.0000318742	-0.000004
37.0000	0.0000617547	0.0000406851	0.0000359688	-0.000004
38.0000	0.0000730774	0.0000476779	0.0000423986	-0.000004
39.0000	0.0000830700	0.0000540435	0.0000486227	-0.000004
40.0000	0.0000901469	0.0000588070	0.0000517504	-0.000004
41.0000	0.0000953110	0.0000622014	0.0000541636	-0.000004
42.0000	0.0000991074	0.0000645162	0.0000563568	-0.000004
43.0000	0.0001029710	0.0000670357	0.0000585840	-0.000004
44.0000	0.0001060591	0.0000690565	0.0000604468	-0.000004
45.0000	0.0001090590	0.0000709690	0.0000620078	-0.000004
46.0000	0.0001120632	0.0000727919	0.0000637651	-0.000004
47.0000	0.0001152433	0.0000746921	0.0000657357	-0.000004
48.0000	0.0001181696	0.0000765475	0.0000679010	-0.000004
49.0000	0.0001212229	0.0000785003	0.0000697566	-0.000004
50.0000	0.0001242888	0.0000803130	0.0000721697	-0.000004
51.0000	0.0001279027	0.0000823139	0.0000744212	-0.000004
52.0000	0.0001312018	0.0000845594	0.0000767982	-0.000004
53.0000	0.0001352057	0.0000868814	0.0000785114	-0.000004

54.0000	0.0001392111	0.0000868723	0.0000825613	-0.00000010
55.0000	0.0001431934	0.0000908279	0.0000865959	-0.00000008
56.0000	0.0001475364	0.0000935064	0.0000896391	-0.00000007
57.0000	0.0001520836	0.0000959528	0.0000931007	-0.00000005
58.0000	0.0001556557	0.0000984862	0.0000967255	-0.00000003
59.0000	0.0001599696	0.0001008645	0.0001006619	-0.00000002
60.0000	0.0001636242	0.0001036360	0.0001055916	-0.00000001
61.0000	0.0001674493	0.0001063234	0.0001093911	0.00000000
62.0000	0.0001712726	0.0001089515	0.0001139939	0.00000000
63.0000	0.0001759144	0.0001116251	0.0001184010	0.00000000
64.0000	0.0001804812	0.0001139289	0.0001228280	0.00000000
65.0000	0.0001854171	0.0001168312	0.0001273081	0.00000000
66.0000	0.0001897237	0.0001196681	0.0001305438	0.00000001
67.0000	0.0001943767	0.0001236307	0.0001347541	0.00000001
68.0000	0.0001991257	0.0001266354	0.0001385009	0.00000002
69.0000	0.0002042394	0.0001302695	0.0001423528	0.00000003
70.0000	0.0002096224	0.0001335875	0.0001464086	0.00000004
71.0000	0.0002152158	0.0001377795	0.0001513642	0.00000005
72.0000	0.0002201937	0.0001406788	0.0001556020	0.00000006
73.0000	0.0002254053	0.0001440686	0.0001594999	0.00000007
74.0000	0.0002316638	0.0001486598	0.0001646039	0.00000008
75.0000	0.0002371676	0.0001524146	0.0001689261	0.00000008
76.0000	0.0002431331	0.0001556680	0.0001732660	0.00000009
77.0000	0.0002497201	0.0001599423	0.0001781289	0.00000010
78.0000	0.0002564164	0.0001643577	0.0001834165	0.00000010
79.0000	0.0002632263	0.0001691096	0.0001890660	0.00000011
80.0000	0.0002728368	0.0001749128	0.0001963395	0.00000012
81.0000	0.0002826490	0.0001810220	0.0002033284	0.00000012
82.0000	0.0002964873	0.0001911310	0.0002111578	0.00000013
83.0000	0.0003144502	0.0002033676	0.0002199315	0.00000015
84.0000	0.0003390945	0.0002190669	0.0002306007	0.00000016
85.0000	0.0003764977	0.0002450238	0.0002500727	0.00000018
86.0000	0.0004330993	0.0002629960	0.0002807505	0.00000022
87.0000	0.0006022413	0.0003958667	0.0003704334	0.00000040
88.0000	0.0004746531	0.0005829828	0.0005154652	0.00000075
89.0000	0.0010024015	0.0006668678	0.0005838092	0.00000094
90.0000	0.0010721035	0.0007140977	0.0006263086	0.00000106
91.0000	0.0011265214	0.0007506026	0.0006576739	0.00000116
92.0000	0.0011778153	0.0007856230	0.0006861812	0.00000125
93.0000	0.0012234695	0.0008155349	0.0007110817	0.00000132
94.0000	0.0012607755	0.0008414289	0.0007358133	0.00000140
95.0000	0.0013030332	0.0008704799	0.0007671032	0.00000150
96.0000	0.0013433116	0.0008952050	0.0007932837	0.00000160
97.0000	0.0014258666	0.0009515447	0.0008353190	0.00000172
98.0000	0.0015240221	0.0010589150	0.0009172775	0.00000191
99.0000	0.0016085926	0.0012068250	0.0010362556	0.00000225

DATA ANALYSIS

Bogen/Smith Feb. 1986

COMPOUNDS: VCL, VOCL, DCE, BNZ, CLFH

STATIONS A,B,D, and F: Weighted Average Individual Risk

Weights:	A .25	A .25
	B .25	E .40
	D .25	C .15
	F .25	F .20

Percentile

Percentile

0.	0.	C.	0.0000
1.0000	0.0000119460	0.00001324E3	1.0000
2.0000	0.0000156075	0.0000164325	2.0000
3.0000	0.0000179668	0.0000188E21	3.0000
4.0000	0.0000207219	0.0000208173	4.0000
5.0000	0.0000223633	0.00002332E7	5.0000
6.0000	0.0000254948	0.0000257504	6.0000
7.0000	0.0000278966	0.0000282544	7.0000
8.0000	0.0000300740	0.0000308486	8.0000
9.0000	0.0000322668	0.0000336372	9.0000
10.0000	0.0000337775	0.0000356674	10.0000
11.0000	0.0000359101	0.0000379173	11.0000
12.0000	0.0000375922	0.0000397931	12.0000
13.0000	0.0000396047	0.0000413343	13.0000
14.0000	0.0000412700	0.0000429528	14.0000
15.0000	0.0000427910	0.0000444416	15.0000
16.0000	0.0000443832	0.0000459243	16.0000
17.0000	0.0000459137	0.0000477889	17.0000
18.0000	0.0000475331	0.0000492297	18.0000
19.0000	0.0000491222	0.0000507661	19.0000
20.0000	0.0000505528	0.0000523307	20.0000
21.0000	0.0000521641	0.0000538351	21.0000
22.0000	0.0000534101	0.0000553665	22.0000
23.0000	0.0000551213	0.0000570069	23.0000
24.0000	0.0000567850	0.0000584052	24.0000
25.0000	0.0000583104	0.0000600574	25.0000
26.0000	0.0000595291	0.0000615665	26.0000
27.0000	0.0000609398	0.0000632521	27.0000
28.0000	0.0000624920	0.0000647120	28.0000
29.0000	0.0000639084	0.0000663750	29.0000
30.0000	0.0000652984	0.0000679663	30.0000
31.0000	0.0000666952	0.0000690913	31.0000
32.0000	0.0000681751	0.0000706720	32.0000
33.0000	0.0000695121	0.0000717976	33.0000
34.0000	0.0000708344	0.0000733306	34.0000
35.0000	0.0000721014	0.0000748485	35.0000
36.0000	0.0000733940	0.0000763210	36.0000
37.0000	0.0000746910	0.0000778514	37.0000
38.0000	0.0000763045	0.0000793699	38.0000
39.0000	0.0000777944	0.0000809702	39.0000
40.0000	0.0000792022	0.0000826075	40.0000
41.0000	0.0000809779	0.0000841434	41.0000
42.0000	0.0000826667	0.0000856116	42.0000
43.0000	0.0000843548	0.0000874245	43.0000
44.0000	0.0000860280	0.0000892629	44.0000
45.0000	0.0000875945	0.0000909096	45.0000
46.0000	0.0000892188	0.0000927660	46.0000
47.0000	0.0000912793	0.0000945362	47.0000
48.0000	0.0000933251	0.0000963741	48.0000

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49.0000	0.00004188	0.0000980939	49.0000
50.0000	0.0000475545	0.0000999289	50.0000
51.0000	0.0000996262	0.0001021408	51.0000
52.0000	0.0001016491	0.0001046748	52.0000
53.0000	0.0001039926	0.0001071460	53.0000
54.0000	0.0001064349	0.0001097920	54.0000
55.0000	0.0001096245	0.0001126752	55.0000
56.0000	0.0001122482	0.0001150656	56.0000
57.0000	0.0001157567	0.0001176408	57.0000
58.0000	0.0001186060	0.0001207582	58.0000
59.0000	0.0001227197	0.0001234717	59.0000
60.0000	0.0001267467	0.0001271274	60.0000
61.0000	0.0001316333	0.0001309652	61.0000
62.0000	0.0001369030	0.0001349603	62.0000
63.0000	0.0001434428	0.0001390285	63.0000
64.0000	0.0001501922	0.0001438627	64.0000
65.0000	0.0001591538	0.0001485198	65.0000
66.0000	0.0001710379	0.0001543214	66.0000
67.0000	0.0001816358	0.0001592774	67.0000
68.0000	0.0001896090	0.0001641432	68.0000
69.0000	0.0001989361	0.0001704458	69.0000
70.0000	0.0002064036	0.0001776602	70.0000
71.0000	0.0002137381	0.0001852956	71.0000
72.0000	0.0002190300	0.0001941856	72.0000
73.0000	0.0002266648	0.0002053501	73.0000
74.0000	0.0002326950	0.0002179657	74.0000
75.0000	0.0002394111	0.0002308161	75.0000
76.0000	0.0002450625	0.0002499023	76.0000
77.0000	0.0002520022	0.0002657587	77.0000
78.0000	0.0002596123	0.0002859498	78.0000
79.0000	0.0002673615	0.0002999518	79.0000
80.0000	0.0002743790	0.0003112532	80.0000
81.0000	0.0002819232	0.0003220499	81.0000
82.0000	0.0002903360	0.0003327035	82.0000
83.0000	0.0002976524	0.0003409699	83.0000
84.0000	0.0003068053	0.0003483519	84.0000
85.0000	0.0003149464	0.0003574431	85.0000
86.0000	0.0003249839	0.0003657447	86.0000
87.0000	0.0003346745	0.0003751234	87.0000
88.0000	0.0003439268	0.0003841729	88.0000
89.0000	0.0003536567	0.0003944676	89.0000
90.0000	0.0003653216	0.0004060516	90.0000
91.0000	0.0003775376	0.0004197684	91.0000
92.0000	0.0003887416	0.0004345696	92.0000
93.0000	0.0004076259	0.0004516638	93.0000
94.0000	0.0004266342	0.0004724605	94.0000
95.0000	0.0004508652	0.0004960444	95.0000
96.0000	0.0004789169	0.0005232969	96.0000
97.0000	0.0005128299	0.0005591788	97.0000
98.0000	0.0005533487	0.0006160105	98.0000
99.0000	0.0006155258	0.0006927988	99.0000

COMPOUNDS: VCL, PCE, TCE, VOCL, CCE, BNZ, CLFM

Calculated Distributions of Individual Risk

Percentile	STATION A	STATION B	STATION C	STATION F
0.	0.	0.	C.	-0.00004410C
1.0000	0.0000167416	0.0000104513	0.00001055e7	-0.000038111
2.0000	0.0000201954	0.00001255E1	0.0000124066	-0.000032112
3.0000	0.0000225259	0.0000139631	0.0000137900	-0.00002681E
4.0000	0.0000242155	0.0000142549	0.0000147735	-0.000021675
5.0000	0.0000258961	0.000015e690	0.0000156949	-0.000017248
6.0000	0.0000271932	0.0000164566	0.0000165174	-0.00001311e
7.0000	0.0000283096	0.0000171433	0.0000172849	-0.00000861e
8.0000	0.0000295966	0.000017E3e1	0.000018C432	-0.000004945
9.0000	0.0000307031	0.0000184297	0.0000187433	-0.000003748
10.0000	0.0000316542	0.0000190747	0.0000192151	-0.000003170
11.0000	0.0000326966	0.0000197551	0.0000199449	-0.000002666
12.0000	0.0000338039	0.0000202387	0.0000205676	-0.000002277
13.0000	0.0000346373	0.0000207854	0.0000210972	-0.000001950
14.0000	0.0000355356	0.0000213115	0.0000217091	-0.000001620
15.0000	0.0000363311	0.000021E004	0.00002232E0	-0.000001340
16.0000	0.0000370161	0.0000222389	0.0000227976	-0.000001050
17.0000	0.0000378112	0.0000227703	0.0000233639	-0.000000720
18.0000	0.0000385936	0.0000232621	0.0000238953	-0.000000430
19.0000	0.0000393749	0.000023e722	0.0000244075	-0.000000160
20.0000	0.0000401131	0.0000241711	0.0000249112	0.000000100
21.0000	0.0000409161	0.000024e785	0.0000253937	0.000000037
22.0000	0.0000418316	0.0000252247	0.0000258708	0.000000065
23.0000	0.0000428108	0.000025E257	0.00002650E0	0.000000091
24.0000	0.0000438087	0.0000265473	0.0000272412	0.000000116
25.0000	0.0000445363	0.000027C8E5	0.000027E9e2	0.000000140
26.0000	0.0000454770	0.0000277467	0.0000285023	0.000000163
27.0000	0.0000464649	0.0000284356	0.0000291974	0.000000187
28.0000	0.0000475088	0.0000291322	0.000029E928	0.000000210
29.0000	0.0000486140	0.000029E595	0.00003054E9	0.000000236
30.0000	0.0000497405	0.0000305784	0.0000313912	0.000000259
31.0000	0.0000513301	0.0000315562	0.0000323316	0.000000282
32.0000	0.0000527962	0.000032e300	0.0000333016	0.000000310
33.0000	0.0000545322	0.000033E427	0.0000345954	0.000000331
34.0000	0.0000565647	0.0000351800	0.0000357623	0.000000361
35.0000	0.0000596773	0.000036E202	0.00003759e6	0.000000384
36.0000	0.0000643914	0.0000397141	0.0000402446	0.000000409
37.0000	0.0000760322	0.0000464039	0.00004478e1	0.000000439
38.0000	0.0000915892	0.00005798E8	0.0000522974	0.000000462
39.0000	0.0000990680	0.00006432E1	0.000057E410	0.000000494
40.0000	0.0001044989	0.0000684469	0.0000e13447	0.000000520
41.0000	0.000109009E	0.0000715235	0.0000e43258	0.000000552
42.0000	0.0001135631	0.0000741705	0.000067C2E1	0.000000582
43.0000	0.0001166602	0.0000767352	0.00006911E8	0.000000615
44.0000	0.0001206058	0.0000792412	0.0000715299	0.000000652
45.0000	0.0001241229	0.0000CF1e2E1	0.00007415e5	0.00000068E
46.0000	0.0001284115	0.0000844175	0.0000769191	0.000000734
47.0000	0.0001323431	0.0000875e23	0.00008029E5	0.000000778
48.0000	0.0001368215	0.00009C39E3	0.0000835524	0.000000840
49.0000	0.0001420116	0.0000937007	0.000087E839	0.000000900
50.0000	0.0001484406	0.0000980230	0.0000929224	0.000000950
51.0000	0.0001554625	0.0001027172	0.00009F7470	0.000001068
52.0000	0.0001631518	0.000106E5e6	0.000105E147	0.000001140
53.0000	0.0001714261	0.0001117505	0.000115371E	0.000001230

54.0000	0.0001 3300	0.0001151393	C.00011254001	C.00000131966
55.0000	0.0001881400	0.0001193971	C.0001351193	0.00000140539
56.0000	0.0001967505	0.0001240129	C.0001430806	0.00000148532
57.0000	0.0002053309	0.0001291605	C.0001507263	0.00000155704
58.0000	0.0002127309	0.0001334420	C.0001572401	0.00000163612
59.0000	0.0002211341	0.0001382014	C.0001637301	0.00000170979
60.0000	0.0002291074	0.0001430089	C.0001705682	0.00000179092
61.0000	0.0002374263	0.0001474257	C.0001769954	0.00000186926
62.0000	0.0002460355	0.0001521943	C.0001830809	0.00000196638
63.0000	0.0002532971	0.0001565996	C.0001893402	0.00000207714
64.0000	0.0002604108	0.0001605321	C.0001961033	0.00000220924
65.0000	0.0002685594	0.0001647855	C.0002022566	0.00000236661
66.0000	0.0002754990	0.0001685263	C.0002082651	0.00000258823
67.0000	0.0002823051	0.0001726795	C.0002145035	0.00000282531
68.0000	0.0002890742	0.0001766196	C.00022196703	0.00000313247
69.0000	0.0002954087	0.0001810707	C.0002293651	0.00000354304
70.0000	0.0003035968	0.0001850113	C.0002309116	0.00000386657
71.0000	0.0003120290	0.0001894472	C.0002371761	0.00000421539
72.0000	0.0003188752	0.0001930911	C.0002419357	0.00000454371
73.0000	0.0003255634	0.0001975205	C.0002471573	0.00000487792
74.0000	0.0003337023	0.0002025943	C.0002536896	0.00000523578
75.0000	0.0003422117	0.0002076196	C.0002599254	0.00000554247
76.0000	0.0003491150	0.0002118553	C.0002654939	0.00000586652
77.0000	0.0003567085	0.0002168571	C.0002716460	0.00000620299
78.0000	0.0003672591	0.0002220305	C.0002784427	0.00000655260
79.0000	0.0003757695	0.0002278940	C.0002854810	0.00000689765
80.0000	0.0003869332	0.0002345229	C.0002941267	0.00000731710
81.0000	0.0003968601	0.0002425033	C.0003031442	0.00000773982
82.0000	0.0004160306	0.0002519356	C.0003141768	0.00000823501
83.0000	0.0004327652	0.0002636519	C.0003244769	0.000008661481
84.0000	0.0004527254	0.0002782673	C.0003377407	0.00000905731
85.0000	0.0004891809	0.0003004160	C.0003560954	0.00000960447
86.0000	0.0005543724	0.0003352321	C.0003845051	0.00001031195
87.0000	0.0007449347	0.0004478697	C.0004589469	0.00001117394
88.0000	0.0009677429	0.0006302672	C.0005836126	0.00001230126
89.0000	0.0010927630	0.0007125543	C.0006546236	0.00001365960
90.0000	0.0011622346	0.0007621924	C.0007014063	0.00001511233
91.0000	0.0012161805	0.0008010866	C.0007320404	0.00001658916
92.0000	0.0012656278	0.0008343418	C.0007656641	0.00001797781
93.0000	0.0013107589	0.0008637734	C.0007958769	0.00001928594
94.0000	0.0013481090	0.0008919118	0.0008218055	0.00002043543
95.0000	0.0013977059	0.0009262662	C.0008550131	0.00002171679
96.0000	0.0014431083	0.0009553564	C.0008847672	0.00002277545
97.0000	0.0015204839	0.0010077254	C.0009340050	0.00002429122
98.0000	0.0015660416	0.0011076273	C.0010001422	0.00002586257
99.0000	0.0018914242	0.0012574432	C.0011262981	0.00002881467

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COMPOUNDS: VCL, PCE, TCE, VOCL, DCE, BNZ, CLFM

STATIONS A,B,C, and F: Weighted Average Individual Risk

Percentile	Weights:				(test)			
	A	B	C	F	A	B	C	F
0.	0.	0.	0.	0.	0.	0.	0.	0.
1.0000	0.0000204584	0.0000242386	0.0000279106	0.0000311852	0.0000304999	0.0000341969	0.0000381523	0.0000413655
2.0000	0.0000349727	0.0000378378	0.0000408607	0.0000437934	0.0000467985	0.0000497234	0.0000524977	0.0000551247
3.0000	0.0000524977	0.0000551247	0.0000582948	0.0000605146	0.0000633737	0.0000656121	0.0000674726	0.0000695475
4.0000	0.0000633737	0.0000656121	0.0000674726	0.0000695475	0.0000719375	0.0000744198	0.0000770200	0.0000794324
5.0000	0.0000719375	0.0000744198	0.0000770200	0.0000794324	0.0000816143	0.0000837869	0.0000857691	0.0000876149
6.0000	0.0000816143	0.0000837869	0.0000857691	0.0000876149	0.0000894127	0.0000914198	0.0000935049	0.0000953721
7.0000	0.0000894127	0.0000914198	0.0000935049	0.0000953721	0.0000977037	0.0000997545	0.0001019075	0.0001037152
8.0000	0.0000977037	0.0000997545	0.0001019075	0.0001037152	0.0001058274	0.0001078090	0.0001097040	0.0001117766
9.0000	0.0001058274	0.0001078090	0.0001097040	0.0001117766	0.0001136784	0.0001158186	0.0001180233	0.0001203513
10.0000	0.0001136784	0.0001158186	0.0001180233	0.0001203513	0.0001229854	0.0001252108	0.0001275451	0.0001297198
11.0000	0.0001229854	0.0001252108	0.0001275451	0.0001297198	0.0001325839	0.0001353098	0.0001381520	0.0001410089
12.0000	0.0001325839	0.0001353098	0.0001381520	0.0001410089	0.0001438811	0.0001467611	0.0001496489	0.0001525445
13.0000	0.0001438811	0.0001467611	0.0001496489	0.0001525445	0.0001554479	0.0001583589	0.0001612774	0.0001642035
14.0000	0.0001554479	0.0001583589	0.0001612774	0.0001642035	0.0001671381	0.0001700799	0.0001730289	0.0001760851
15.0000	0.0001671381	0.0001700799	0.0001730289	0.0001760851	0.0001791479	0.0001822181	0.0001852957	0.0001883797
16.0000	0.0001791479	0.0001822181	0.0001852957	0.0001883797	0.0001914689	0.0001945655	0.0001976695	0.0002007809
17.0000	0.0001914689	0.0001945655	0.0001976695	0.0002007809	0.0002038989	0.0002070245	0.0002101577	0.0002132985
18.0000	0.0002038989	0.0002070245	0.0002101577	0.0002132985	0.0002164369	0.0002195829	0.0002227365	0.0002258977
19.0000	0.0002164369	0.0002195829	0.0002227365	0.0002258977	0.0002290559	0.0002322215	0.0002353945	0.0002385749
20.0000	0.0002290559	0.0002322215	0.0002353945	0.0002385749	0.0002417489	0.0002449305	0.0002481197	0.0002513165
21.0000	0.0002417489	0.0002449305	0.0002481197	0.0002513165	0.0002545109	0.0002577129	0.0002609225	0.0002641397
22.0000	0.0002545109	0.0002577129	0.0002609225	0.0002641397	0.0002673439	0.0002705559	0.0002737757	0.0002769935
23.0000	0.0002673439	0.0002705559	0.0002737757	0.0002769935	0.0002802099	0.0002834339	0.0002866557	0.0002898755
24.0000	0.0002802099	0.0002834339	0.0002866557	0.0002898755	0.0002930939	0.0002963199	0.0002995435	0.0003027649
25.0000	0.0002930939	0.0002963199	0.0002995435	0.0003027649	0.0003059839	0.0003092099	0.0003124335	0.0003156549
26.0000	0.0003059839	0.0003092099	0.0003124335	0.0003156549	0.0003188779	0.0003221019	0.0003253235	0.0003285439
27.0000	0.0003188779	0.0003221019	0.0003253235	0.0003285439	0.0003317639	0.0003349859	0.0003382089	0.0003414299
28.0000	0.0003317639	0.0003349859	0.0003382089	0.0003414299	0.0003446489	0.0003478699	0.0003510889	0.0003543059
29.0000	0.0003446489	0.0003478699	0.0003510889	0.0003543059	0.0003575239	0.0003607439	0.0003639619	0.0003671779
30.0000	0.0003575239	0.0003607439	0.0003639619	0.0003671779	0.0003703939	0.0003736119	0.0003768279	0.0003800419
31.0000	0.0003703939	0.0003736119	0.0003768279	0.0003800419	0.0003832579	0.0003864739	0.0003896879	0.0003928999
32.0000	0.0003832579	0.0003864739	0.0003896879	0.0003928999	0.0003961139	0.0003993279	0.0004025399	0.0004057509
33.0000	0.0003961139	0.0003993279	0.0004025399	0.0004057509	0.0004089619	0.0004121719	0.0004153809	0.0004185879
34.0000	0.0004089619	0.0004121719	0.0004153809	0.0004185879	0.0004217939	0.0004250019	0.0004282079	0.0004314119
35.0000	0.0004217939	0.0004250019	0.0004282079	0.0004314119	0.0004346179	0.0004378239	0.0004410279	0.0004442309
36.0000	0.0004346179	0.0004378239	0.0004410279	0.0004442309	0.0004474339	0.0004506369	0.0004538379	0.0004570379
37.0000	0.0004474339	0.0004506369	0.0004538379	0.0004570379	0.0004602399	0.0004634419	0.0004666419	0.0004698409
38.0000	0.0004602399	0.0004634419	0.0004666419	0.0004698409	0.0004730399	0.0004762379	0.0004794339	0.0004826279
39.0000	0.0004730399	0.0004762379	0.0004794339	0.0004826279	0.0004858239	0.0004890179	0.0004922109	0.0004954019
40.0000	0.0004858239	0.0004890179	0.0004922109	0.0004954019	0.0004985939	0.0005017839	0.0005049719	0.0005081579
41.0000	0.0004985939	0.0005017839	0.0005049719	0.0005081579	0.0005113439	0.0005145279	0.0005177109	0.0005208919
42.0000	0.0005113439	0.0005145279	0.0005177109	0.0005208919	0.0005240739	0.0005272539	0.0005304319	0.0005336079
43.0000	0.0005240739	0.0005272539	0.0005304319	0.0005336079	0.0005367839	0.0005399579	0.0005431299	0.0005462999
44.0000	0.0005367839	0.0005399579	0.0005431299	0.0005462999	0.0005494739	0.0005526439	0.0005558119	0.0005589779
45.0000	0.0005494739	0.0005526439	0.0005558119	0.0005589779	0.0005621439	0.0005653079	0.0005684699	0.0005716299
46.0000	0.0005621439	0.0005653079	0.0005684699	0.0005716299	0.0005747939	0.0005779539	0.0005811119	0.0005842679
47.0000	0.0005747939	0.0005779539	0.0005811119	0.0005842679	0.0005874239	0.0005905779	0.0005937299	0.0005968799
48.0000	0.0005874239	0.0005905779	0.0005937299	0.0005968799	0.0005999839	0.0006030839	0.0006061809	0.0006092759

49.0000	0.0001378552	C.0001377227	.0000793015	C.00005048
50.0000	0.0001407559	C.0001400774	C.0000813439	0.00005207
51.0000	0.0001435693	C.0001427931	C.0000835487	C.00005369
52.0000	0.0001464524	C.0001452183	C.0000854501	0.00005505
53.0000	0.0001493753	C.0001483283	C.0000880061	C.00005654
54.0000	0.0001523748	C.0001508127	C.0000907375	0.00005837
55.0000	0.0001555732	C.0001540627	C.0000928356	0.00005978
56.0000	0.0001598618	0.0001575702	C.0000951295	C.00006128
57.0000	0.0001643265	0.0001612272	C.0000975491	C.00006280
58.0000	0.0001689505	C.0001644625	C.0001002750	C.00006447
59.0000	0.0001730946	C.0001684695	C.0001028761	0.00006627
60.0000	0.0001776716	0.0001723213	C.0001055364	0.00006757
61.0000	0.0001836000	C.0001764688	C.0001078213	C.00006954
62.0000	0.0001892681	0.0001806079	C.0001102704	0.00007118
63.0000	0.0001947449	0.0001853479	C.0001130905	0.00007294
64.0000	0.0002017037	0.0001895807	C.0001158380	C.00007457
65.0000	0.0002095510	C.0001941297	C.0001186579	C.00007663
66.0000	0.0002180812	C.0001996267	C.0001215133	0.00007835
67.0000	0.0002247096	0.0002056536	C.0001243996	0.00008000
68.0000	0.0002307941	0.0002109181	C.0001274813	C.00008164
69.0000	0.0002384554	0.0002172988	C.0001306388	0.00008368
70.0000	0.0002457089	0.0002248247	C.0001334368	0.00008568
71.0000	0.0002528513	0.0002332646	C.0001363476	0.00008778
72.0000	0.0002599996	C.0002420739	C.0001394772	0.00008950
73.0000	0.0002673251	0.0002552182	C.0001428066	0.00009257
74.0000	0.0002743479	0.0002667731	C.0001464428	C.00009477
75.0000	0.0002805328	0.0002807976	C.0001499200	0.00009731
76.0000	0.0002883072	0.0002967754	C.0001533030	0.00010009
77.0000	0.0002966641	0.0003118931	C.0001564810	0.00010310
78.0000	0.0003052506	0.0003275154	C.0001593584	0.00010581
79.0000	0.0003129182	C.0003374013	C.0001628595	0.00010894
80.0000	0.0003204447	0.0003495175	C.0001660885	0.00011245
81.0000	0.0003272209	C.0003627074	C.0001695919	0.00011657
82.0000	0.0003350463	0.0003710674	C.0001732141	0.00012060
83.0000	0.0003432383	0.0003807889	C.0001781006	0.00012479
84.0000	0.0003520433	C.0003905885	C.0001830383	0.00013019
85.0000	0.0003614785	0.0004001783	C.0001869070	0.00013558
86.0000	0.0003710391	C.0004098882	C.0001923480	C.00014159
87.0000	0.0003811896	0.0004183790	C.0001960053	0.00014761
88.0000	0.0003913370	0.0004295794	C.0002011726	C.00015437
89.0000	0.0004007554	0.0004421347	C.0002060741	0.00016099
90.0000	0.0004137709	0.0004544519	C.0002132248	C.00016692
91.0000	0.0004266316	0.0004664131	C.0002214344	0.00017822
92.0000	0.0004428630	0.0004833953	C.0002288343	0.00018665
93.0000	0.0004584003	0.0004998501	C.0002368908	0.00019627
94.0000	0.0004791230	C.0005183414	C.0002455478	C.00020701
95.0000	0.0005015589	C.0005425868	C.0002553466	C.00021811
96.0000	0.0005291854	C.0005745288	C.0002656395	C.00023137
97.0000	0.0005669994	0.0006124660	C.0002801838	0.00024618
98.0000	0.0006076190	0.0006667789	C.0003014875	C.00026437
99.0000	0.0006669510	0.0007551034	C.0003365391	C.00029707

Table 2.3
Calculation of Increased Annual Carcinogenic Risk
to Individuals for Comparative Purposes

Compound	EPA 95% UCL Potency ^a (mg/kg/day) ⁻¹	Individual Dose ^b (mg/kg/day)	Annual Risk ^c ($\times 10^6$)
Vinyl Chloride	.0175	2.5×10^{-4}	.062
Perchloroethylene	.051	2.0×10^{-4}	.15
Trichloroethylene	.011	2.3×10^{-4}	.036
Vinylidene Chloride	1.18	1.1×10^{-4}	1.8
1,2-Dichloroethane	.091	1.9×10^{-4}	.25
Benzene	.029	9.4×10^{-5}	.039
Chloroform	.007	1.4×10^{-5}	.014
Total =			2.4

^a EPA potency values are for lifetime exposure and are taken from the EPA's Health Assessment Document for Trichloroethylene (EPA 600/8-82/006F, July 1985) p. 9-133.

^b Doses derived from "Mean" values presented in Table 2.1, multiplied by the factor $(15/24) \times (0.5) \times (0.5) = 0.15625$ to account for partial daily exposure, dilution and partial absorption, as explained in the text.

^c Annual risk calculated using 70 years as human lifespan.

Table 2.4
Table of Comparative Cancer Risks

Daily Exposure Situation	Person at Risk ^a	Dose (mg/kg/day)	Individual Annual Risk ($\times 10^6$)
BKK Resident	Adult Youth	See Table 2.3 See Table 2.3	2.4 3.4
Drinking 2 liters of average U.S. tap water containing 88 ppb chloroform	Adult	2.45×10^{-3}	2.5
Swimming for 1 hr in a pool	Child	0.01	10
Eating a peanut butter sandwich containing 2 teaspoons of p. butter @ 2 ppb aflatoxin B ₁	Adult Child	0.914×10^{-6} 2.133×10^{-6}	38 88
Worker exposed to 50 ppm perchloroethylene for 8 hr. e.g., in a dry cleaning shop, for 5 days/week	Adult	32.3	~ 810,000
Worker exposed to 5 ppm vinylidene chloride for 8 hr. e.g., in an adhesive manufacturing plant. for 5 days/week	Adult	1.90	~ 890,000

^a Exposure Assumptions: Adult weighs 70 kg and inhales 20 m³ of air per day; Youth are aged 0-15 breathing a time-weighted average of 0.4143 m³/kg/day (equivalent to 29 m³ per 70 kg per day); and Child weighs 30 kg.

APPENDIX E

ANALYSIS OF VARIANCE OF THE DATA IN EPA AND DHS' 1984 EXTENDED MONITORING PROGRAM

Analysis of variance (ANOVA) is a statistical technique that isolates and assesses the contribution of one or more factors to an outcome of interest. Independent categorical variables (factors) are investigated for their influence on the dependent continuous random variable (outcome). Assumptions underlying ANOVA are: independence of the values, normality of the errors of the values, and equal variance for all observations. Two types of models used with ANOVA are the fixed effects model and the mixed effects model. In the fixed effects model there is one level of effect for each group of observations, and the assumption is that the difference is due to the fixed effect. In the random effects model, the effects are drawn from a prior distribution in which it is not possible to estimate the magnitude of the effects for one group of observations, but it is possible to estimate the overall variance. The mixed effects model contains both fixed and random effects.

For the data in EPA and DHS' 1984 Extended Monitoring Program (CH2MHill, 1988), the independent categorical variables were: "type" (Priority I homes/control home), "agent" (chemicals detected in the Priority I homes), and "sample date/calendar period"; the dependent random variable was "concentrations detected".

Three-way ANOVA was performed on the data in the program for the primary factor of interest, "type" (that is, whether or not there was a difference between the concentrations detected in the Priority I homes and the concentrations detected in the control home). The two other factors investigated to ensure that the difference was not due to their influence were "agent" (vinyl chloride, 1,1-dichloroethylene, 1,2-dichloroethane, 1,1,1-trichloroethane, perchloroethylene, trichloroethylene, and benzene) and "sample date". A mixed effects model was used, with agent and date as random effects. Analysis required log transformation of the values recorded in Volume 5 Appendix H-2. The table below shows that, in 3-way ANOVA, concentrations detected in the Priority I homes were significantly higher than concentrations detected in the control home ($p < 0.039$).

3-WAY MIXED EFFECTS ANALYSIS OF VARIANCE

PARAMETER	ESTIMATE	STANDARD DEVIATION	P VALUE
Error Variance	0.049	0.003	-
Constant	0.354	0.106	0.001
Type	0.053	0.026	0.039
Agent	0.072	0.039	-
Date	0.009	0.004	-

A 2-way fixed effects model ANOVA was also done to verify that the sample date did not affect the primary factor of interest, "type". The analysis was done separately for the early part of the month (calendar period October 1-14) and the later part of the month (calendar period October 15-29). The second factor in the analysis was "agent" (the six compounds detected in the Priority I homes). Benzene was omitted in the 2-way ANOVA because too few data points in the control home precluded its inclusion in the analysis. Values recorded in Volume 5 Appendix H-2 were log transformed for the analysis. The table below shows that, in 2-way ANOVA, concentrations detected in the Priority I homes were significantly higher than concentrations detected in the control home ($p < 0.006$).

2-WAY FIXED EFFECTS ANALYSIS OF VARIANCE
FOR ALL DATES COMBINED

SOURCE	SUM OF SQUARES	DEGREES OF FREEDOM	MEAN SQUARE	F VALUE	P VALUE
Agent	4.6203	5	0.9241	15.13	0.0000
Type	0.4713	1	0.4713	7.72	0.0057
Interaction	0.5490	5	0.1098	1.80	0.1117
Error	28.2087	462	0.0611		