

Occupational exposures and pancreatic cancer: a meta-analysis

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Abstract

Objectives—Consolidation of epidemiological data on pancreatic cancer and worksite exposures.

Methods—Publications during 1969–98 were surveyed. Studies without verified exposures were excluded. Meta-analyses were conducted on data from 92 studies covering 161 populations, with results for 23 agents or groups of agents. With a standard format, five epidemiologists extracted risk estimates and variables of the structure and quality of each study. The extracted data were centrally checked. Random meta-models were applied.

Results—Based on 20 populations, exposure to chlorinated hydrocarbon (CHC) solvents and related compounds was associated with a meta-risk ratio (MRR) of 1.4 (95% confidence interval (95% CI) 1.0 to 1.8). Nickel and nickel compounds were considered in four populations (1.9; 1.2 to 3.2). Excesses were found also for chromium and chromium compounds (1.4; 0.9 to 2.3), polycyclic aromatic hydrocarbons (PAHs) (1.5; 0.9 to 2.9), organochlorine insecticides (1.5; 0.6 to 3.7), silica dust (1.4; 0.9 to 2.0), and aliphatic and alicyclic hydrocarbon solvents (1.3; 0.8 to 2.8). Evidence on pancreatic carcinogenicity was weak or non-positive for the following agents: acrylonitrile (1.1; 0.0 to 6.2); arsenic (1.0; 0.6 to 1.5); asbestos (1.1; 0.9 to 1.5); diesel engine exhaust (1.0; 0.9 to 1.3); electromagnetic fields (1.1; 0.8 to 1.4); formaldehyde (0.8; 0.5 to 1.0); flour dust (1.1; 0.3 to 3.2); cadmium and cadmium compounds (0.7; 0.4 to 1.4); gasoline (1.0; 0.8 to 1.2); herbicides (1.0; 0.8 to 1.3); iron and iron compounds (1.3; 0.7 to 2.5); lead and lead compounds (1.1; 0.8 to 1.5); man-made vitreous fibres (1.0; 0.6 to 1.6); oil mist (0.9; 0.8 to 1.0); and wood dust (1.1; 0.9 to 2.5). The occupational aetiological fraction of pancreatic cancer was estimated at 12%. In a subpopulation exposed to CHC solvents and related compounds, it was 29%; to chromium and chromium compounds, 23%; to nickel and nickel compounds, 47%; to insecticides, 33%; and to PAHs, 33%.

Conclusion—Occupational exposures may increase risk of pancreatic cancer. High quality studies are called for on interactions between occupational, environmental, and lifestyle factors as well as

interactions between genes and the environment.

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Keywords: pancreatic cancer; occupational exposure; meta-analysis

Some 180 000 pancreatic cancers are registered annually in the world. It is highly and rapidly fatal and represents the fifth leading cause of deaths from cancer in industrialised countries and is 50%–100% more common in men than in women. It is not consistently associated with socioeconomic status within national populations, although there is a tendency toward higher age adjusted risk in richer than poorer countries.^{1–3} Incidence has risen in industrialised countries since the 1960s and subsequently levelled off in several populations.⁴

The causes of pancreatic cancer are mostly unknown. Tobacco smoking is the single established common cause. The proportion of cases attributable to smoking has been estimated at 5%–50%, depending on the population.⁵ Epidemiology of pancreatic cancer has suffered from bias due to high misclassification rate, notably prominent when case definition has been based only on death certificates. This has resulted in inconsistencies in the results on the aetiological role of environmental and occupational dererminants of pancreatic cancer.^{6–8}

We identified published epidemiological studies about pancreatic cancer and job titles, industrial branches, and occupational exposures, and conducted a meta-analysis of the role of 33 chemical or physical agents present in the working environment that affect the aetiology of pancreatic cancer.

Materials and methods

The literature search covered the Medline, Toxline, and Cancerlit databases for the period 1969 to May 1998, with the following search conditions:

- (1) (occupational OR agriculture) AND neoplasms AND morbidity
- (2) (occupational OR agriculture) AND neoplasms AND mortality NOT morbidity
- (3) (occupational OR agriculture) AND neoplasms AND incidence NOT mortality NOT morbidity
- (4) (pancreatic OR digestive) AND occupational
- (5) (pancreatic OR digestive) AND case AND (control OR referent)

The search was accompanied by a scan of the lists of reference of the identified studies. In all, 1902 studies were identified. A total of 365 studies remained after exclusion of studies that did not report on pancreatic cancer; that did not represent the most recent update; that reponed insufficient data for the meta-analysis; that did not report data for any job or occupational agent; that did not report original results (reviews); that reported on part of a larger population reported elsewhere; and that reponed on job categories or agent categories too broad or outside our list of job titles and agents. The agents were based on the FINJEM job exposure matrix.⁹ The list of job titles covered 150 entries in the Finnish social status categories 3, 4, and 5. Data for categories 1 and 2 represented the highest social categories and were excluded because the relevant occupational chemical and physical exposures were minimal or nonexistent. The chemical and physical agents considered were the following 23 agents or groups of agents: aliphatic and alicyclic hydrocarbon solvents; aromatic hydrocarbon solvents (excluding aromatic amines); arsenic; asbestos; cadmium and cadmium compounds; chlorinated hydrocarbon (CHC) solvents and related compounds (excluding organochlorine insecticides); chromium and chromium compounds; diesel engine exhaust; electromagnetic fields, flour dust, formaldehyde, fungicides, gasoline, herbicides, insecticides, iron and iron compounds; lead and lead compounds; man made vitreous fibres; nickel and nickel compounds; oil mist (including machining fluid and cutting fluid); polycyclic aromatic hydrocarbons (PAH); silica dust; and wood dust.

The studies were divided into (a) agent specific studies with direct risk estimates for one or several of the 23 agents, or for job titles with verified exposure(s) to the agent(s), and (b) job specific studies without risk estimates for any of the selected agents but instead for one or more of the 150 job categories without verified exposure(s) to the agent(s). This report is based on the data set from agent specific studies only (92 studies presenting data for 161 different exposed populations).⁸⁻⁹⁹

Standardised data extraction forms (available from the corresponding author) covered characteristics of the study (publication year, country, study type, case definition, source of cases, reference population, selection of control subjects, follow up period, loss to follow up, response rates, sources of exposure or job data, time reference for exposure or job; risk measure, cohort admission, lag periods, exposure-response, job title coding applied), risk estimates, latency periods, and numbers of exposed cases.

Five epidemiologists (AO, TP, NJ, EW, CW) read the reports and extracted the necessary data, using predefined rules. The main principles of extraction were to extract:

- Relevant, unbiased estimates of relative risk
- Measures of relative risks associated with specific exposures and job titles

- The most unbiased estimate if there is a choice and choose:
- Estimates adjusted for at least known risk factors for pancreatic cancer (age, sex, tobacco smoking), if there was a choice
- Social class adjusted risk ratios over those unadjusted for social class
- Risk estimate nearest to 20 y latency period, if there is a choice.

The extracted data were then centrally checked for consistency (AO, TP) and finally entered into a database and checked for correctness.

We recovered missing 95% CIs with Byar's approximation¹⁰⁰ in cohort studies, and with variance of log odds ratio in case-control studies.

Simple random effects models¹⁰¹ were applied in estimating the meta-risk ratios (MRR). Fixed effects models were used only on occasion for Comparison with results from random effects models. Standard errors of log of

Table 1 Characterisation of the 92 agent specific studies

	Populations (n)
Study type:	
Administrative (linkage of administrative records or PMR/PCMR MOR studies)	23
Industrial cohort	88
Industry based (nested) case-control study	7
Population or hospital based case-control study	43
Cases:	
Exocrine pancreatic cancers only	32
All pancreatic cancers	127
Unspecified	2
Diagnosis of cases:	
Histological	47
Other (clinical: radiology, necropsy, etc)	2
Mortality files	96
Mixed	9
Unknown	7
Ascertainment of cases:	
Mortality files	98
Cancer registry files	40
Hospital records	21
Mixed	1
Unspecified	1
sex:	
Men	112
Women	8
Both men and women or unspecified	41
Risk measure:	
SMR (standardised mortality ratio)	68
SIR (standardised incidence ratio)	17
PMR (proportional mortality ratio)	18
MOR (mortality odds ratio)	1
HR (hazard ratio)	3
OR (odds ratio)	50
RR (risk ratio)	4
Source of exposure data:	
Industrial hygiene measurements	4
Job exposure matrix	15
Expert assessment	25
Job titles†	57
Other‡	19
Mixed	37
Unknown	4
Time reference of exposure:	
Last or around diagnosis	17
Earlier cross section	9
Lifetime longitudinal	47
Less than lifetime longitudinal	80
Other	1
Unknown	1
Total	161

[†]Ratio of risks or cumulative incidences in exposed and unexposed cohorts

[‡]Job titles with specifically verified exposure(s) to the relevant agent(s)

[§]Employment duration, biological monitoring, employer record, registry of chemicals, self reporting

risk ratios (RRs), when not given, were recovered as the log of the ratio of upper and lower 95% CIs divided by 3.92. A test of the heterogeneity was performed as a χ^2 test with degrees of freedom equal to one less than the number of populations. The public domain software package *computer programs for epidemiologic analysis* (PEPI; <http://www.usd-inc.com/pepi.html>), version 3.0, and the statistical software package *Stata* release 5 were used.¹⁰²

Population aetiological fractions were estimated as $PEF = p - p/MRR$, and aetiological fractions among the exposed groups as $EEF = 1 - 1/MRR$ for each agent where excesses were observed. The proportion of pancreatic cancer cases who were exposed (p), was calculated as $\Sigma \text{exposed cases} / \Sigma \text{all cases}$, summing over studies that provided the necessary numbers. The PEFs were based on MRRs from population based and hospital based case-control studies only, whereas the EEFs were based on MRRs from all except proportional studies. Aetiological fractions were not calculated for agents with no indication of increased risk.

The data were organised and analysed by populations rather than studies. This was because most studies considered more than one subpopulation defined by exposure. A total of 161 populations were covered in agent specific studies (table 1). Most (85) populations were from North America, closely followed by Western Europe (63). There were few populations from Central and Eastern Europe, Oceania, and none from Middle and South America, Asia, or Africa. The annual number of studies has been rising considerably during 1969–8 (two studies during 1969–79; 58 studies during 1980–9, and 101 studies during 1990–8).

Industrial cohort (88 populations) and case-control (50 populations) studies were the most common agent specific studies (table 1). The cases represented predominantly all pancreatic

cancers, irrespective of type. As most studies considered mortality, the diagnosis was in most populations obtained from the death record. Less than 5% of the populations were women. Exposures were assessed in 57 populations through job titles; in 25 through expert assessments; in 15 through job exposure matrices (JEMs); and in 60 through other, mixed, or unexplained methods. Industrial hygiene measurements were explicitly applied in four populations only. Exposure assessment was longitudinal for 127 populations.

Reference populations in cohort studies represented predominantly national or other large populations (57 studies). Fewer studies (23) used local populations as the reference; even fewer (seven) used an internal reference. Follow up for case ascertainment began usually during 1940–79 (99 studies) and spanned in most studies a period of ≥ 10 years (98 studies). Losses to follow up in agent specific cohort studies were minor or moderate ($< 5\%$ in 44 studies but unknown in 51 studies) but were more marked in job branch cohort studies. Agent specific cohorts were rather evenly distributed between entry cohorts, cross sectional cohorts, and mixed cohorts.

Case-control studies used variable periods of case ascertainment. Most agent specific case-control studies (44) used cancer or population controls. Response rates were $\leq 80\%$ for the cases, and $\leq 90\%$ in the controls. Sixteen studies did not report response rates.

Results

The aggregated results for the occupational agents are shown in table 2. Random effects models without covariates were used. Proportional studies (four) representing 18 populations, were excluded from most analyses because of poor quality. The study by Magnani *et al.*,⁸² although reported as a case-control study, used job and branch data as well as diagnoses of pancreatic cancer from death certificates only. We therefore treated it as a proportional mortality rate (PMR) study.

Significant excesses were found for nickel and nickel compounds (MRR 1.9; 1.2 to 3.9; four populations) and CHC solvents and related compounds (1.4; 95% CI 1.0 to 1.8; 20 populations). Non-significant excesses over MRR > 1.3 were found for PAHs; organochlorine insecticides; silica dust; and chromium and chromium compounds.

Table 3 shows the MRRs after stratification by sex and diagnostic quality. Only seven studies presented results for women. For CHC solvents and related compounds, the MRR for women (1.8; 95% CI 0.7 to 5.8; three populations) was higher than for men (1.3; 0.9 to 1.9; 14 populations). The MRR remained essentially unchanged irrespective of whether histological verification of diagnoses of pancreatic cancer was done or not. For chromium and chromium compounds and for PAHs, the excess risks disappeared in studies that had histological diagnoses, whereas for nickel and nickel compounds, the MRR was increased in populations with and without histological verification of the cases.

Table 2 Populations (n), meta-risk estimates (MRRs) (95% CIs), ranges of point estimates and p values for heterogeneity by occupational agents (proportional studies excluded: simple random effects models with no covariates)

Agent	n	MRR	95% CI	Range of point estimate	p Value
Aliphatic and alicyclic hydrocarbon solvents	3	1.3	0.8 to 2.0	1.0–1.6	0.09
Aromatic hydrocarbon solvents	13	1.0	0.8 to 1.3	0.5–2.9	0.2
Arsenic	4	1.9	0.6 to 1.6	0.9–1.4	0.9
Asbestos	24	1.1	0.9 to 1.4	0.5–3.6	0.004
Cadmium and cadmium compounds	2	0.7	0.4 to 1.4	0.7–4.8	0.9
Chlorinated hydrocarbon solvents and related compounds	20	1.4	1.0 to 1.8	0.3–4.9	0.05
Chromium and chromium compounds	9	1.4	0.9 to 2.3	0.6–20	0.2
Diesel engine exhaust	7	1.0	0.9 to 1.2	0.5–1.4	0.2
Electromagnetic fields	5	1.1	0.8 to 1.4	0.6–2.4	0.04
Flour dust	1	1.1	0.3 to 3.2		
Formaldehyde	5	0.8	0.5 to 1.0	0.5–1.0	0.3
Fungicides	2	1.3	0.4 to 3.8	1.3–1.4	0.9
Gasoline	1	1.0	0.8 to 1.3	0.2–1.1	0.5
Herbicides	10	1.0	0.8 to 1.3	0.6–5.9	0.3
Insecticides	3	1.5	0.6 to 3.7	0.8–21.0	0.1
Iron and iron compounds	1	1.3	0.7 to 2.5		
Lead and lead compounds	4	1.1	0.8 to 1.5	1.0–1.4	1.0
Man made vitreous fibres	5	1.0	0.0 to 1.6	0.3–6.8	0.03
Nickel and nickel compounds	4	1.9	1.2 to 3.2	1.2–3.6	0.9
Oil mist	1	0.9	0.8 to 1.0	0.3–1.0	0.5
Polycyclic aromatic hydrocarbons (PAHs)	4	1.5	0.9 to 2.5	1.3–3.6	0.8
Silica dust	3	1.4	0.9 to 2.0	1.1–2.0	0.2
Wood dust	4	1.1	0.8 to 1.5	0.7–1.1	0.3

Table 3 Populations (n), metarisk estimates (MRRs) (95% CIs) by occupational agents, sex, and quality of diagnosis: (proportional studies are excluded; simple random effects models with covariates)

Agent	Sex						Histological diagnosis								
	Men			Women			Unspecified or both			Yes			No		
	n	MRR	95% CI	n	MRR	95% CI	n	MRR	95% CI	n	MRR	95% CI	n	MRR	95% CI
Aliphatic and alicyclic hydrocarbon solvents										2	1.3	0.8 to 2.0			
Aromatic hydrocarbon solvents	6	0.7	0.6 to 1.0				7	1.3	0.9 to 1.7	4	1.0	0.6 to 1.8	9	1.0	0.7 to 1.2
Arsenic	3	0.9	0.6 to 1.6							2	1.1	0.5 to 2.7	2	0.9	0.5 to 1.6
Asbestos	18	1.3	1.0 to 1.5	3	0.8	0.4 to 1.7	3	0.6	0.4 to 0.8	3	0.7	0.4 to 1.2	21	1.2	1.0 to 1.5
Chlorinated hydrocarbon solvents and related compounds	14	1.3	0.9 to 1.9	3	1.8	0.7 to 4.6	3	1.6	0.8 to 2.9	4	1.2	0.7 to 2.1	16	1.4	1.0 to 2.1
Chromium and chromium compounds	7	1.8	0.9 to 3.6				2	0.9	0.5 to 1.8	3	1.0	0.7 to 1.6	6	2.3	0.9 to 5.8
Diesel engine exhaust	5	1.1	0.9 to 1.3				2	0.9	0.6 to 0.9	3	1.1	0.9 to 1.2	4	1.0	0.9 to 1.1
Electromagnetic fields	6	1.1	0.9 to 1.3										6	1.1	0.9 to 1.3
Formaldehyde	3	0.8	0.5 to 1.3				2	0.6	0.3 to 1.1	2	0.5	0.3 to 0.9	3	0.9	0.7 to 1.3
Gasoline	2	1.0	0.8 to 1.3							2	1.0	0.8 to 1.4	2	0.9	0.6 to 1.3
Herbicides	8	1.2	0.8 to 2.0				2	0.9	0.7 to 1.2				9	1.1	0.8 to 1.5
Insecticides	2	0.7	0.4 to 1.5				2	3.7	0.3 to 43.3				3	1.2	0.3 to 4.3
Lead and lead compounds	3	1.1	0.8 to 1.6							2	1.1	0.8 to 1.6	2	1.0	0.5 to 2.2
Man made vitreous fibres	2	1.8	0.8 to 3.8				3	0.9	0.7 to 1.2	3	0.8	0.5 to 1.2	2	1.2	0.8 to 1.6
Nickel and nickel compounds	3	2.0	1.2 to 3.5							2	2.0	1.2 to 3.2	2	1.6	0.4 to 6.9
Oil mist	4	0.9	0.8 to 1.0				2	0.5	0.3 to 1.1				6	0.9	0.7 to 1.2
Polycyclic aromatic hydrocarbons (PAHs)	3	1.8	0.8 to 3.4							2	1.4	0.8 to 2.3	2	3.0	0.7 to 13.2
Silica dust	2	1.1	0.8 to 1.5							2	1.5	0.8 to 2.6			
Wood dust							3	1.1	0.7 to 1.7	3	1.2	0.9 to 1.6			

Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMWSIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0).

Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMWSIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results

Table 4 Populations (n), metarisk ratios (MRRs) (95% CIs), by study type (proportional studies excluded; simple random effects models with no covariates)

Agent	Case-control studies and cohort studies with internal reference			SMR/SIR studies		
	n	MRR	95% CI	n	MRR	95% CI
Aliphatic and alicyclic hydrocarbon solvents	2	1.3	0.8 to 2.0			
Aromatic hydrocarbon solvents	4	0.9	0.5 to 1.6	9	1.0	0.8 to 1.3
Arsenic	3	1.2	0.5 to 2.6	1	0.9	0.4 to 1.5
Asbestos	4	0.7	0.5 to 1.0	20	1.2	1.0 to 1.5
Cadmium and cadmium compounds	1	0.8	0.2 to 2.9	1	0.7	0.3 to 1.4
Chlorinated hydrocarbon solvents and related compounds	4	1.4	0.8 to 2.4	16	1.3	0.9 to 2.0
Chromium and chromium compounds	3	1.0	0.7 to 1.6	6	2.3	0.9 to 5.8
Diesel engine exhaust	5	1.1	0.9 to 1.4	2	0.9	0.8 to 1.1
Electromagnetic fields				5	1.1	0.8 to 1.4
Flour dust	1	1.1	0.3 to 3.2			
Formaldehyde	2	0.5	0.3 to 1.6	3	0.9	0.7 to 1.3
Fungicides	2	1.3	0.3 to 3.8			
Gasoline	2	1.0	0.8 to 1.4	2	0.9	0.6 to 1.3
Herbicides	1	0.9	0.7 to 1.8	9	1.0	0.8 to 1.3
Insecticides	2	3.7	0.3 to 43.3	1	0.8	0.3 to 1.7
Iron and iron compounds	1	1.3	0.7 to 2.5			
Lead and lead compounds	3	1.1	0.8 to 1.6	1	1.0	0.4 to 2.1
Man made vitreous fibres	4	0.9	0.4 to 2.2	1	1.1	0.8 to 1.5
Nickel and nickel compounds	2	2.0	1.2 to 3.2	2	1.6	0.4 to 6.9
Oil mist	3	0.8	0.6 to 1.3	3	0.9	0.7 to 1.0
Polycyclic aromatic hydrocarbons (PAHs)	2	1.4	0.8 to 2.3	2	3.0	0.7 to 13.2
Silica dust	2	1.5	0.8 to 2.7	1	1.2	0.6 to 2.1
Wood dust	4	1.2	0.9 to 1.6			

Table 5 Populations (n), population aetiological fractions (PEFs), aetiological fractions among exposed (EEFs), (95% CIs)

Agent	n	Population aetiological fraction (PEF)					Aetiological fraction among exposed (EEF)				
		Proportion exposed	MRR	95% CI	PEF	95% CI	n	MRR	95% CI	EEF	95% CI
Aliphatic and alicyclic hydrocarbon solvents	2	0.094	1.3	0.6 to 2.0	0.022	0.00 to 0.047	2	1.3	0.8 to 2.0	0.23	0.00 to 0.50
Chlorinated hydrocarbon solvents and related compounds	3	0.061	1.2	0.7 to 1.9	0.010	0.00 to 0.029	20	1.4	1.0 to 1.8	0.29	0.00 to 0.44
Chromium and chromium compounds	3	0.034	1.1	0.7 to 1.6	0.003	0.00 to 0.013	9	1.4	0.9 to 2.3	0.29	0.00 to 0.57
Fungicides	1	0.0034	1.4	0.3 to 7.2	0.001	0.00 to 0.0029	2	1.3	0.4 to 3.8	0.23	0.00 to 0.74
Insecticides							3	1.5	0.6 to 3.7	0.33	0.00 to 0.73
Iron and iron compounds	1	0.067	1.3	0.7 to 2.5	0.016	0.00 to 0.040	1	1.3	0.7 to 2.5	0.23	0.00 to 0.60
Lead and lead compounds	2	0.11	1.1	0.8 to 1.6	0.010	0.00 to 0.041	4	1.1	0.8 to 1.5	0.09	0.00 to 0.33
Nickel and nickel compounds	2	0.063	2.0	1.2 to 3.2	0.032	0.011 to 0.043	4	1.9	1.2 to 3.2	0.47	0.17 to 0.69
Polycyclic aromatic hydrocarbons (PAHs)	2	0.029	1.4	0.6 to 2.3	0.008	0.00 to 0.016	4	1.5	0.9 to 2.4	0.33	0.00 to 0.58
Silica dust	2	0.063	1.5	0.8 to 2.7	0.021	0.00 to 0.040	3	1.4	0.9 to 2.0	0.29	0.00 to 0.50
Wood dust	4	0.062	1.2	0.9 to 1.6	0.01	0.00 to 0.023	5	1.2	0.9 to 1.6	0.17	0.00 to 0.38

Only agents with both MRRs >1.0 considered. Negative lower confidence bounds of PEFs and EEFs forced to zero.

MRR=meta risk ratio, as estimated from population based and hospital based case-control studies for PEF, and all but proportional studies for EEF

concerned organochlorine compounds. There were 12 further studies that considered unspecified pesticides. This group of agents was considered to be too heterogeneous in exposures and was dropped from the analysis.

Given the low proportions of the populations that were exposed, the PEFs remained low, from 0.1%–3% (table 5). Assuming independence between exposures, summing up of the PEFs resulted in an overall aetiological fraction of 12% for workplace exposures. The agent specific EEFs ranged from 9% to 47%, with wide 95% CIs. In a subpopulation exposed to nickel and nickel compounds, the EEF was 47% (95% CI 17 to 69%); to PAHs, 33% (0 to 58%); to insecticides, 33% (0 to 73%); to CHC solvents and related compounds, 29% (0 to 44%); to chromium and chromium compounds, 29% (0 to 57%); and to aliphatic and alicyclic solvents, 23% (0 to 50%).

Discussion

EVIDENCE

The excess risk found for CHC solvents and related compounds was based on 20 populations. Heterogeneity of RRs was nearly significant and may be explained by differences in the quality and exposure level of the agents. Various compounds with variable carcinogenic potential were mentioned as worker exposures: trichloroethylene, tetrachloroethylene, 1,1,1-trichloroethane, methylene chloride, vinyl chloride, ethylene chlorohydrine, ethylene dichloride, bis(chloromethyl)ether, and polychlorinated biphenyls. Intensities and long term doses were characterised in most of the studies either poorly or not at all.

The risk for nickel and nickel compounds was most evident in population based case-control studies. For chromium and chromium compounds, the MRR was non-significantly increased in all studies, but was not in excess in population based case-control studies. For PAHs, a non-significant increase was present in all studies, in population-based case-control studies, and in the two SMR/SIR studies. These findings could have occurred by chance.

The excess of silica dust reached significance in one³² of three studies. This same Finnish population-based case-control study found a significant excess for aliphatic and dicyclic hydrocarbon solvents. This finding was aggre-

gated with the finding of no excess for alkanes (C₅-C₁₀) in another population based study from Montreal,⁷⁹ the result being an MRR of 1.3 (95% CI 0.8 to 2.0).

Two case-control studies,⁴² both based on cytological diagnoses, one SMR study,⁴³ and one PMR study⁴⁴ considered exposure to organochlorine insecticides. The insecticidal agents listed as exposures were 1,1-bis(4-chlorophenyl)-2,2,2-trichloroethane (DDT), chlordane, heptachlor, endrin, aldrin, dieldrin, bulan, chlorfenethol, chloropropylate, dicofol, ethylan, methoxychlor, and tetrachlorodiphenylethane (TDE). Excluding the PMR study, the aggregated MRR was 1.5 (95% CI 0.6 to 3.7). In case-control studies it was 3.7 (0.3 to 43) based on the random effects model, and 1.7 (0.9 to 3.2) based on the fixed effects model (heterogeneity p=0.09). The highest RR was obtained for exposure to the DDT family (DDT, ethylan, DDD; OR 21.0; 95% CI 2.6 to 966; five exposed cases) in a case-control study nested in a chemical manufacturing cohort.⁴⁵ Potential confounders included nitrophenol derivatives, clays, N,N-dimethylformamide, dispersing agents, octane, and carbon tetrachloride. The other case-control study⁴⁶ was population based (Michigan, US), with self reported exposures. Based on 21 exposed cases, it yielded an OR of 1.5 (95% CI 0.8 to 2.9) for organochlorine insecticides. Assuming an effect, the difference between the two point estimates might be due to qualitative and quantitative differences in exposures between manufacturing and agricultural application.

There is a possibility of effect modification of environmental or occupational determinants by lifestyles (tobacco, alcohol, coffee) or dietary factors. These interactions were not considered in the studies. Also, genetic factors may interact with environmental or occupational exposures.

COMBINABILITY AND HETEROGENEITY

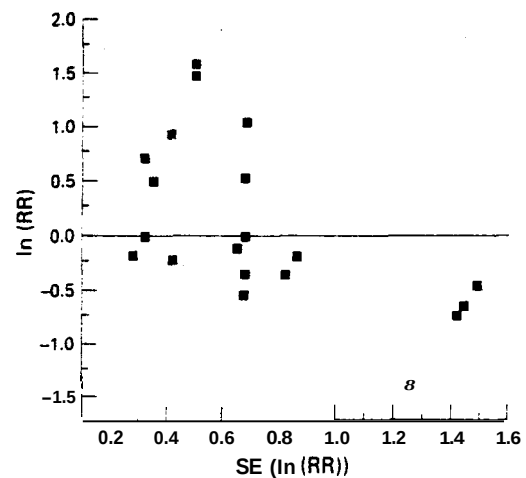
Epidemiological meta-analyses have imperfect combinability of results associated with different study types, methods, populations, exposure circumstances, and diagnostic specificities. We calculated MRRs including and excluding the poorest quality studies (proportional studies). We also calculated separate MRRs for cohort studies with internal controls.

case-control studies, SMR/SIR studies, and the few proportional studies. Differences in results from different study types were not consistent.

Several populations were poorly characterised. There were even studies that did not specify whether the cohort consisted of men, women, or both. We analysed the data for known male and female populations separately, and found data for women to be associated with a slightly higher *MRR* than data for men for CHC solvents and related compounds.

There was in all likelihood substantial heterogeneity across populations in the quality and intensity of exposure categories, in the intake route (respiratory, dermal, or other) of exposure, time aspects of exposure (period, latency, duration, quality, and intensity), and applied scales of exposure, as well as in the quality of diagnosis of pancreatic cancer. Qualitative and quantitative differences in exposures have already been exemplified in connection with CHC solvents and related compounds and organochlorine insecticides. Based on a rough statistical test, significant heterogeneity of risk was found for asbestos, electromagnetic fields, CHC solvents and related compounds, and man made vitreous fibres. Some studies did not document the exposure aspects at all, and no study provided a full documentation. Expert assessments, which represent an imperfect yet acceptable method of exposure assessment, were used in 25 populations. Industrial hygiene measurements that represent a certain degree of objectivity were used as the prime source of exposure data in only four populations. Exposure assessment based on job titles (57 populations) are of lower quality, unless the exposures are highly homogeneous within job titles. However, some of the populations represented rather homogeneous single title cohorts. Job exposure matrices assess exposures better if the matrix is specific for branch and job title, even for company and period. The JEMs of variable degree of specificity were applied in 15 populations. Most were relatively unspecific and thereby induced exposure misclassification. Misclassification, however, was likely to be non-differential, resulting in underestimation of the *MRR*. Multiple sources of exposure data were applied in 37 populations. Exposure data were longitudinal in 127 populations and lifelong in 47 populations. The longitudinality of exposures was thus well covered.

Misclassification rates for pancreatic cancer are marked, particularly when the diagnosis is based on death certificate only. This adds to the bias in the meta estimates towards the null value. Garabrandt *et al.*² showed what the magnitude of the impact may be in a single study. They compared, in a case-control study of pancreatic cancer, ORs for the DDT family between cases representing death certificates and cases representing cytological verification. For death certificate cases the ORs ranged from 0.8 to 2.6; for cytologically verified cases, from 15.4 to infinity. In our meta-analysis, agent specific *MRRs* were higher for nine agents but lower for 10 agents in populations in



Funnel plot for chlorinated hydrocarbon solvents. Natural log of risk ratio ($\ln(RR)$) plotted against imprecision (SE of $\ln(RR)$). The unit of the plot is study population.

which histological verification was applied, compared with no histological verification. For CHC solvents and related compounds, the two *MRRs* were practically identical. For PAHs and chromium, the *MRRs* were higher in populations with histological diagnoses.

PUBLICATION BIAS

Publication bias is not likely in this study, as very few small studies expressly considered the occupational determinants of pancreatic cancer. Non-positive occupational findings from small studies therefore were not likely to remain unpublished. Also, a funnel plot for CHC solvents and related compounds, for which an excess risk was found (figure), did not identify a concentration of small studies (high *SE* of $\ln(RR)$) at high *RRs*. Publication bias therefore may be claimed to be minimal or non-existent. Negative results based on reasonably large numbers are unlikely to remain unpublished.

A counterargument may be raised about cohort studies with multiple end points. Some of these studies deleted results based on small numbers, occasionally for pancreatic cancer. This omission may have minor influence on the metaresults. Some case-control studies may have omitted results for rare exposures, with similar minor effect on the metaresults.

SELECTION OF STUDIES

We used major databases and lists of references of the studies for the identification of studies. Studies not found in major databases are probably of lower quality. For the same reason and because of the cost we did not try to identify unpublished studies.

EXTRACTION

Extractor bias was minimised by the formal extraction procedure between the extractors, and the central checking of the extraction. The procedure was also intended to guarantee the extraction of the relevant risk estimates in studies that offered several alternative risk estimates.

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